

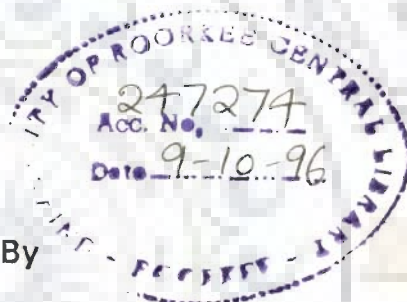
# EFFECT OF WATER QUALITY PARAMETERS ON BIOTA

A THESIS

submitted in fulfilment of the  
requirements for the award of the degree  
of  
DOCTOR OF PHILOSOPHY

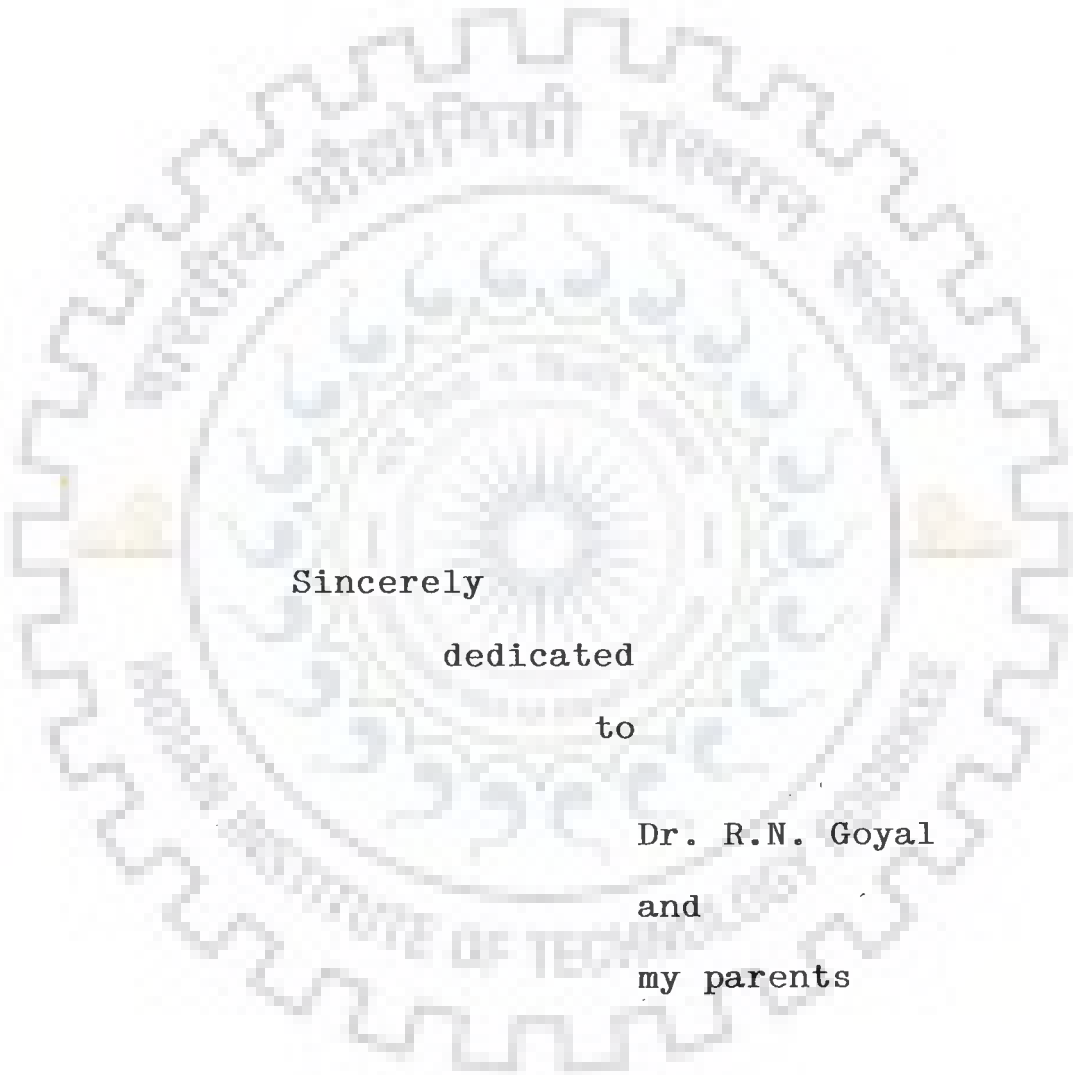
By

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NOVEMBER, 1991



Sincerely

dedicated

to

Dr. R.N. Goyal

and


my parents

CANDIDATE'S DECLARATION

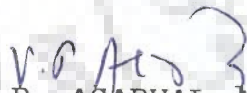
I hereby certify that the work which is being presented in the thesis entitled, 'EFFECT OF WATER QUALITY PARAMETERS ON BIOTA' in fulfilment of the requirement for the award of the degree of Doctor of Philosophy, submitted in the Department of Biosciences and Biotechnology of the University is an authentic record of my own work carried out during the period from February 1989 to November 1991 under the supervision of Dr. R.N. Goyal, Dr. G.S. Randhawa and Dr. V.P. Agarwal.

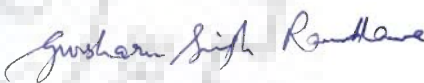
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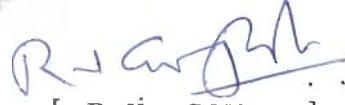
  
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The study of pollution in an ecosystem is an area of global interest particularly as it directly or indirectly affects living beings. One of the major sources of pollution in river ecosystem is water which is polluted by mixing of sub-water stream containing industrial or sewage waste. As the water of river and streams are extensively used for many purposes such as water supply, irrigation and general amenities, it is highly desirable that the sources of water should be kept free from undesirable pollutants so that the interests of the community can be safe guarded.

Saharanpur is an important major industrial city of Western Uttar Pradesh. The waste from numerous industries is mixed in the river Dhamola, which passes through the city. The water of this river is used for agriculture purposes including bathing and drinking by domestic mammals. The water of the river Dhamola just before entering in Saharanpur (North side of city) is colourless and visually looks pure, whereas, after about twelve kms south of entering in Saharanpur, the grey colour of the water can be easily seen. Essential micro and macro nutrients in the river are very much influenced by the mixing of waste discharge. Hence the contamination by sewage and other waste products causes pollution by disturbing the water quality. Apart from this, the sewage has high population of certain pathogenic micro and macro organism which in a rainy season, further affects the

aquatic life and water quality. As the visual changes indicated pollution in the river Dhamola within Saharanpur City, it was considered interesting to study the extent of pollution in this river and is the subject matter of this thesis. The studies include an in-depth analysis of various physico-chemical and biological factors at selected points of the river Dhamola. It has been concluded that the water of this river is not sufficiently pure and hence should not be used for bathing or drinking by human beings and domesticated animals. Attempts have also been made to minimise the pollution load by using certain vascular hydrophytes.

The first chapter of the thesis is General Introduction and highlights the importance of pollution study in present days. The second chapter of the thesis has been devoted to the Literature Review. It reveals the important relevant references reported in the literature relating to the present studies on water pollution. The different methods of water analysis and techniques for reduction of pollutants have been discussed. Biological treatment processes by different species of algae, cyanobacteria and higher vascular hydrophytes and significant results relevant to the present study have also been incorporated at proper places.

The next chapter is devoted to the Materials and Methods. The different methods of analysis of



physico-chemical and biological parameters of river water, industrial and domestic effluents according to the methods specified in "Standard Methods for the Examination of Water and Waste Water" [American Public Health Association, 1987] has been discussed. The details of the experiments for minimising the pollution load of all the samples by using certain vascular hydrophytes (biological treatment), are also presented in this chapter. Three distinct aquatic plant species, viz., Eichhornia crassipes, Lemna minor and Ceratophyllum demorsum were used in stages I to III, whereas in stages IV and V a combination of E. crassipes, L. minor and L. minor, C. demorsum were used. The treatment of the industrial effluent as well as domestic sewage was allowed upto four weeks and the physico-chemical and biological parameters were determined at an interval of one week.

The results obtained on the above studies form the fourth chapter of the thesis. Various parameters determined for the water samples collected at different places of river Dhamola are presented. The samples were collected for a period of two years, i.e., from July 1988 to June 1990 and the results of physico-chemical analysis are presented. The various parameters of the waste water samples collected from sampling stations 4 (domestic sewage) and 5 (industrial effluent) mixing in the river Dhamola clearly indicated that the values of BOD, COD, CO<sub>2</sub> MPN and SPC are much higher than the tolerance limit of

Indian Standard for the discharge of waste water into the inland surface water. Also the samples had a very low value of DO during the entire course of investigation. The water samples obtained from sampling station 5 did not indicate the presence of any planktonic population whereas, in domestic sewage only protozoans were found in all the seasons. Algae was also detected during the rainy season in domestic sewage.

The analysis of Dhamola river water at sampling station 1 indicated the quality of water with the parameters (mg/l) DO [min.7.2], BOD [max.20.5], COD [max.49.0], CO<sub>2</sub> [max.1.1], chloride [max.22.0], total solids [max.489] and total hardness [max.151]. Thus, the water at sampling station 1 was found sufficiently pure for the bathing and agricultural purposes. Due to the discharge of waste water into the river Dhamola, after the entering into the Saharanpur city, the quality of river water changes and the analysis at sampling station 2 (two kms. downstream of the merging point of waste water) indicated it to be polluted. It was found that the values of BOD, COD, CO<sub>2</sub>, total hardness, MFN, SPC, diatoms and protozoans increase tremendously whereas, the values of DO and algae (primary producer) decrease. The values observed for physico-chemical and biological parameters indicated that the water at sampling station 2 was not suitable for bathing purposes due to high value of BOD and coliform bacteria. It's use in bathing may lead to several chronic diseases.

Analysis of water samples obtained from sampling station 3 also indicated the high values for BOD and coliform bacteria. Therefore, it was concluded that the water of the river Dhamola is not suitable for domestic purposes even after seven kms of the mixing of waste water. Other physico-chemical and biological parameters of the water at sampling station 3 are slightly better due to the self purification of river water. The content of the dissolved oxygen in river water reduced significantly at sampling station 2 in comparison to sampling station 1 due to the mixing of organic matter of the waste water. But at station 3 it was found that the concentration of dissolved oxygen increased due to the self purification. The values of different physico-chemical and biological parameters of river water have also been correlated with each other using a package of LOTUS program. An excellent statistical correlation was observed among the following parameters: turbidity - total solids, DO with BOD, COD and algae, BOD with CCD, SPC and algae, COD with SPC and algae and diatoms - protozoans.

Attempts have also been made to use various vasuclar hydrophytes and their combination for the treatment of industrial and domestic effluents to reduce the pollution load of the waste products. Eichhornia crassipes was found most effective in the reduction of pollution load in all the seasons of the year. The values of DO was found to increase in four weeks treatment whereas BOD, CO<sub>2</sub>, COD,

MPN and SPC decreased. The decrease in the value of chloride ions was about 40-50% in comparison to control. As E. crassipes has large leaf area, well developed arenchyma and root systems, more transpiration occurs during photosynthesis and possibly causes an increase in DO value. L. minor used in stage II was found particularly useful for the removal of chloride ( $\text{Cl}^-$ ),  $\text{CO}_2$ , turbidity, total solids and SPC. Hence the combination of E. crassipes and L. minor used in stage IV was found most effective in reducing the pollution load of domestic sewage as well as industrial effluent.

It was observed that before entering in Saharanpur city the water of river Dhamola is pure as indicated by normal values of physico-chemical parameters, but the value of these parameters was found to increase significantly at the mixing point and after mixing point of industrial/domestic discharge into the river. This indicated that a high pollution load has been added in the river in Saharanpur city due to which there is increase in the turbidity, total hardness, total solids, BOD, COD, chloride,  $\text{CO}_2$ , MPN, SPC, protozoans etc.

Thus, the present studies clearly indicate that the river Dhamola becomes polluted after entering into Saharanpur city. The use of plant species, viz., E. crassipes and L. minor can reduce the pollution created by domestic sewage and industrial effluent to some extent and

then water of the river Dhamola will be suitable for bathing purposes. A more vigorous treatment, however, is needed for making it suitable for drinking purposes. It is worth mentioning at this stage that the aquatic plant species used in the present investigation are not only useful for pollution removal but their biomass can also be successfully used for the production of methane gas.



## ACKNOWLEDGEMENT

In the sense of scaling such an laborious task, if for a while I stop and look back, I lost my count to recapitulate the persons and personages who inspired me to accomplish the same.

Eventually it is my great privilege to acknowledge the erudite guidance rendered by Dr. R.N. Goyal, Dr. V.P. Agarwal and Dr. G.S. Randhawa during the course of investigations. I, further extend my deep sense & gratitude to Dr. R.N. Goyal for his intuitive and meticulous guidance, preserving interest, benevolent help and stimulating encouragement. Without his constant encouragement during leave of absence to University of Oklahoma, USA, together with his astute research methodology coupled with assiduity and deep insight into the subject, this work would not have seen the light of the day.

It is my pleasant duty to express my sincere thanks to Prof. S.K. Srivastava (Head, Chemistry Department), Dr. R. Barthwal (Ex-Head of the Biosciences & Biotechnology Department) Prof. C.B. Sharma (Head, Biosciences & Biotechnology Department) and Prof. G.C. Nayak (Ex-Head of the Civil Engineering Department) of the University of Roorkee for providing necessary research facilities.

I wish to express my sincere thanks to Dr. J.K. Grover, Associate professor, Department of Pharmacology, All India Institute of Medical Sciences, New Delhi for providing necessary laboratory facilities.

I greatly acknowledge the help of Soil Conservation Officer, Department of Agriculture, U.P. Government, Saharanpur and District Magistrate, Saharanpur for providing area maps of Dhamola river.

My heartfelt thanks are due to Mr. A.B. Pant and Mr. V.G. Rao (Research Scholars in the Department of Biosciences & Biotechnology) who rendered help in many ways.

Here, I want to make a special mention of Dr. R.N. Goyal who had boosted my moral whenever, I was trapped in the grip of frustration and despair.

Lastly words pale into insignificance in regarding the unflinching love and encouragement rendered by my parents during the course of this investigation by which I attain this level of academic carrier. Really I owe, my parents very much for the same.

DATED : Nov. 10<sup>th</sup>, 1991

Dinesh Kumar Garg  
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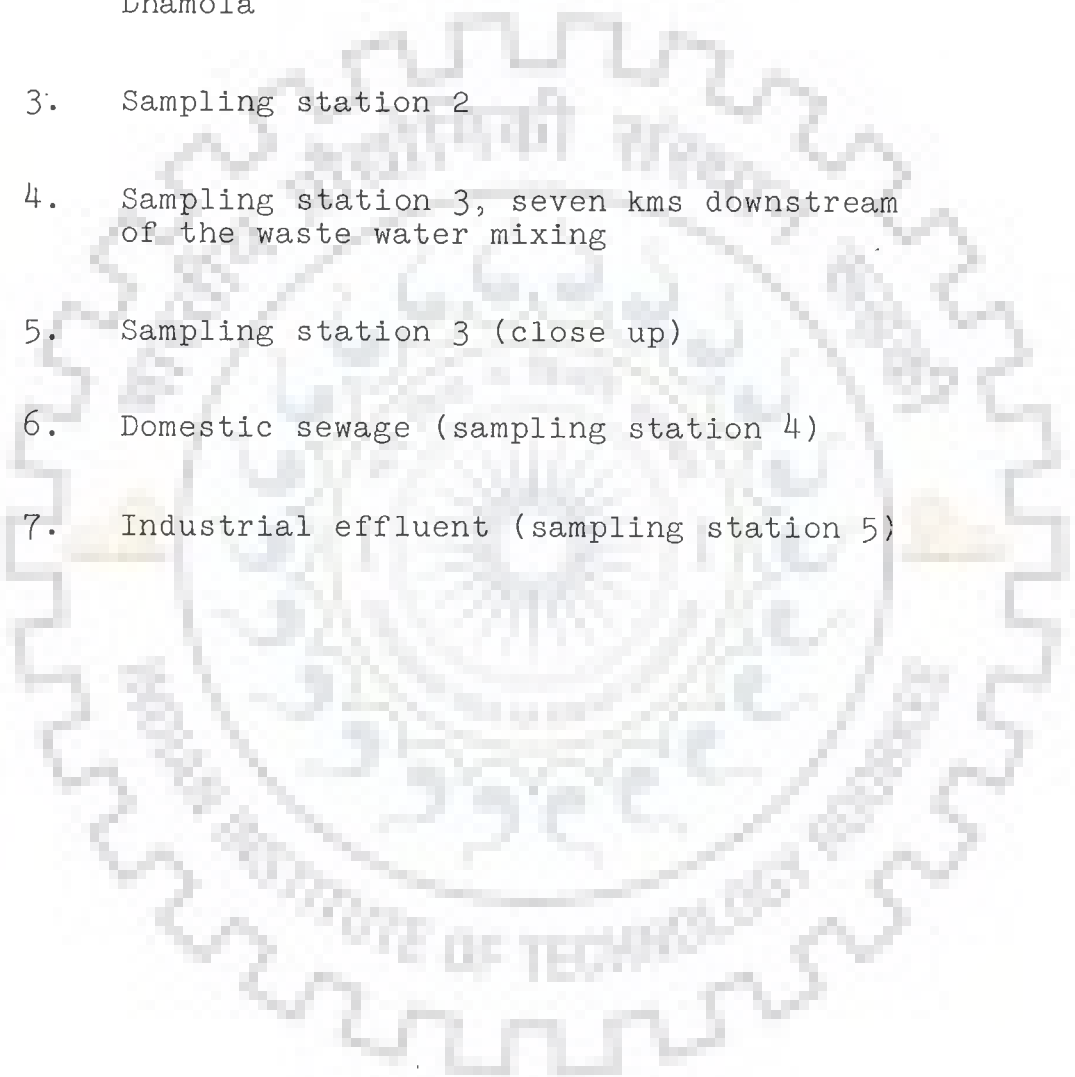
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## LIST OF ABBREVIATIONS USED IN THE THESIS

DO	Dissolved oxygen
BOD	Biochemical oxygen demand
COD	Chemical oxygen demand
MPN	Most probable number
SPC	Standard plate count
ppm	Part per million
Max.	Maximum
Min.	Minimum
S-R	Sedgwick refter
<u>E.crassipes</u>	<u>Eichhornia crassipes</u>
<u>L. minor</u>	<u>Lemna minor</u>
<u>C.demorsum</u>	<u>Ceratophyllum demorsum</u>
U.P.	Uttar Pradesh (state of India)



CHAPTER - I  
INTRODUCTION



The turn of the last two decades has witnessed the development and expansion of research in various fields of pollution study in the entire globe. This period has seen a continued growth in the large scale pollution of air, water and soil due to heavy industrialization and urbanization. The alarming information provided by Union Minister for Environment in the Lok Sabha on August 12, 1991 that Delhi (India) is one of the most polluted cities in the world indicates the importance of monitoring pollution in metropolitan cities. The continuous and rapid increase in population has increased the water requirements for agriculture, industry and for domestic use. Hence water is becoming one of the major constraints for further socio-economic development. As the increase in population presses the need of more food in the world, an efficient planning of water resources for irrigation is a high priority item on the economic development agenda in most countries.

It is well known that, higher productivity and better living standards of a modern society depend on the creation of an infrastructure capable of guaranteeing an adequate supply of water of acceptable quality. Human developments, which were developed in the vicinity of water several decades ago, due to multiple

conveniences provided: water supply, hygiene, water power, transportation, protection, pleasure and scenic beauty has resulted in the large scale pollution of not only water but also of air and soil. The multifaced dependence on water demands strong measures to keep the water free from undesirable pollutants and adaptation of extensive water conservation measures. In this context, it is well known that the rivers and streams are extensively used for many purposes such as water supply, disposal of sewage and trade of effluents, irrigation and land drainage needs, navigation, commercial and recreational fisheries, other forms of recreations and general amenities. Some of these uses create conditions which reduce the capacity of streams to cope with the pollution and strict control measure are necessary, if the interests of the community are to be safeguarded. Conditions in different parts of the country vary widely and the factual information about these conditions is required at regular intervals to asses the current position and the long time effects of the control methods adopted.

### 1.1 POLLUTION AND ITS KINDS

Pollution is an undesirable change in the physical, chemical and biological characteristics of air, water and soil that harmfully affect the animal as well as plant life. The undesirable constituents that pollute

the environment must be controlled for the ecological balance of nature.

In relation to the industrial and domestic waste water, pollution can be divided in four categories:

- a. Physical
- b. Chemical
- c. Physiological
- d. Biological

a. Physical pollution :

Physical pollution consists the changes in the physical characteristics of water bodies after the mixing of industrial as well as domestic waste discharge. These waste discharges may change either of the properties of water bodies, viz., temperature, colour, turbidity, foams, suspended solids and radioactivity etc.

b. Chemical pollution :

Chemical pollution in water is observed when any of the following parameters of water changes, viz., pH, acidity, alkalinity, chlorides, sulphates, sulphides etc. The presence of toxic compounds, heavy metals (biodegradable and non-biodegradable) and organic chemicals in the water bodies also produce chemical pollution. When the pH of water is less than 5.0, it causes corrosion of metals and irritates the skin of

living beings. Both acids and alkalies are toxic to aquatic life and are generally discharged from various industries.

c. Physiological pollution :

Physiological pollution in water is observed due to the unpleasant taste and odour. For example, sulphide, phenols and chloride give bad odour. Iron and manganese salts destroy the taste of water which may harm the aquatic life. Sulphides containing water gives  $H_2S$  and thus smell like rotten eggs.

d. Biological pollution :

The pollution in water caused by some pathogenic micro-organisms, like fungi, algae, protozoans, viruses, bacteria and helminthic parasites which are normally found in domestic waste water is named as biological pollution. It is usually observed where sewage is discharged. These micro-organisms are responsible for many enteric disease like, typhoid, cholera, amoebiasis, tuberculosis, lep-speric, hepatitis, poliomyelitis etc.

## 1.2 WATER POLLUTION PROBLEMS IN INDIA

Since long, the water pollution problems in India have been observed due to the mixing of waste water from various industries as well as domestic waste discharges into the rivers/streams without proper treatment. The pollution in the holy river Ganga is well known as

approximately 270 million litres per day of untreated sullage including 30 million litres per day of industrial effluent was found to be mixed into the river Ganga from Kanpur city alone. In 1987, Maharashtra Pollution Control Board has filed a case against the High Explosive Factory at Khadki near Pune for discharging untreated effluent into the river Mula. Similarly, Central Pollution Control Board, Delhi, has recently reported the increase of pollution load in Yamuna river in Delhi region due to the mixing of various industrial and domestic wastes.

The mixing of this type of wastes discharge into the rivers/streams cause the death of aquatic population like fish and other useful macro-organisms. The discharge of waste water in rivers, e.g. in Kullu river, Chambal, Mahi and Ganga river has seen the tragic loss of fishes.

The various conventional methods used for the disposal of such pollutants as the activated sludge, trickling filters, anaerobic digestions methods etc. are extremely costly from the point of view of the capital financial capacities of local Government authorities. In certain cases raw sewage has been used for the irrigation of crops in sewage farms. Oxidation stabilization ponds (lagoons) which are generally known as low cost alternatives for the treatment of effluents require more operation and maintenance cost as compared to aquatic macrophyte treatment systems.

The use of aquatic hydrophytes for reducing pollution level of industrial effluents and domestic wastes as well as pollution in river water is relatively a new area of research. In recent years, experiments have been conducted in the different laboratories to evaluate the role of hydrophytes in reducing the pollution [1-4]. The plants used in these studies include, Eichhornia crassipes, Lemna minor, Pistia stratiotes, Hydrocotyle umbellata, Azolla caroliniana, Salvinia rotundifolia, Egeria densa etc.

Thus it is clear from the preceding paragraphs that the effluents from industries and municipal waste water affects the water quality, fish population and bottom fauna characteristics of a river [5-10].

The aquatic macrophytic system can be effectively used to reduce the pollution level in the water bodies and the industrial effluents [4]. Physico-chemical and biological studies/surveys of Nagase river in Japan [11], Neva river in USSR [12], Menotre river in Italy [13], Stream of China [14], Big-blue and Sagavonirtok river in USA [15] and Tama river in Japan [16] indicated that the water of these rivers is highly polluted due to mixing of effluents.

### 1.3 SUBJECT MATTER OF THE THESIS

Dhamola is an important river of Western Uttar Pradesh in Saharanpur district, which passes through the

city and runs from North to South. The water of this river is used for agricultural purposes as also for bathing and drinking by domestic animals. The principal source of water in Dhamola is natural water which comes from the hilly region of Northern India (Dehradun hilly region, basin of Himalayas).

The water of this river, just before entering in Saharanpur city (North side of city), is odourless and visually looks pure, whereas after about seven miles (11Kms) south of entering in Saharanpur, the grey colour of the water can be easily seen due to mixing of waste from numerous industries. Essential micro and macro nutrients in the river are very much influenced by the mixing of waste discharge (industrial as well as domestic waste). Hence, the contamination by sewage (domestic waste) and other waste products causes pollution by disturbing the water quality. Apart from this, the sewage, has high population of certain pathogenic micro and macro organisms which in a rainy season further affects the aquatic life and water quality.

Such visual changes indicated pollution in the river Dhamola within Saharanpur city, hence it was considered interesting to study and determine the extent of pollution in this river by industrialization and urbanization of Saharanpur city. Attempts have also been made to cut ways and means to minimize the pollution load of the river Dhamola by using certain vascular

hydrophytes (biological treatment). These studies include an analysis of various physico-chemical and biological factors at selected points of the river Dhamola. It has been found that the water of this river is not sufficiently pure and hence should not be used for bathing, or drinking by human beings and domestic animals.

The present study has been divided into two main parts:

- (A) Analysis of the various physico-chemical and microbiological parameters of the water samples obtained at selected points of the river Dhamola from July 1988 to June 1990. The industrial effluents as well as adjoining city sewage were also analysed before they mix in the river. Statistical analysis of the results is also deduced.
- (B) The role of three hydrophytes viz., Eichhornia crassipes, Lemna minor and Ceratophyllum demorsum in the reduction of pollution load of effluents/waste discharge has also been studied. It has been observed that the use of the combination of two or more plant species in reducing pollution is more efficient in comparison to a single plant.



The thesis has been divided into five chapters and the material presented is as follows:

Chapter	I	Introduction
Chapter	II	Literature Review
Chapter	III	Materials and Methods
Chapter	IV	Results and Discussions
Chapter	V	Bibliography





CHAPTER - II  
LITERATURE REVIEW

The study of pollution in an ecosystem is an area of recent global interest, particularly as it directly or indirectly affects living beings. The increase of population, urbanization and industrialization leads to the generation of large volumes of waste water from domestic, commercial, industrial and other sources [17]. According to Shende and Sunderson [18] about 13,163 million gallons of waste water is released every day from various urban and rural centres of India.

The effluent from chemical industries is one of the most potent source of water pollution. The wastes from industries that are known to cause serious pollution problems, have been attributed by Basu [19] to distillary effluents and pulp and paper mill effluents. The deterioration of water quality of natural water and canal water has been reported due to the mixing of industrial effluents [20].

The disposal of sewage and industrial wastes in land is an age old practice in India. According to Central Board's reports, CUPS/4/1978-79 and CUPS/6/1979-80 [21], out of 142 class-I cities and 190 towns, 52 cities and 43 towns respectively, are disposing their waste water on land [22].

The presence of heavy metals in effluents from the

industries create the phytotoxicity as well as toxicity to human food chain. A comparison of plants irrigated with industrial waste water and domestic sewage indicated that plants irrigated with industrial waste water had high accumulation of heavy metals in comparison to plants irrigated with domestic sewage [23]. The field survey of these areas indicated that the growth of plants in both the cases was suppressed and was not as expected from unpolluted water. The quality of ground water collected from raw sewage irrigated areas was a matter of concern with respect to copper and zinc. The degree of heavy metal contamination was comparatively high in the case of raw sewage irrigated areas whereas copper, lead and cadmium contents were found as 0.6, 0.1 and 0.24 ppm respectively and were more than the permissible range of Central Public Health and Environmental Engineering Organization [24].

In recent years, UNESCO initiated the first multidisciplinary study on the major rivers of the world, which led to interesting and alarming information about the extent of pollution in these rivers. Besides, the well known hydrological features, hydrobiological aspects and primary geochemical cycle operating in different sectors of the rivers were studied in some of the major river of temperate zones in the European continent and North America [25].

Excellent reviews dealing with the ecological aspects of the algae in streams and rivers have been appeared in the literature [26,27]. The phyto plankton and zooplankton studies on number of rivers from Europe and North America have been studied by various workers. These rivers include Thames, Essex, Stour, Lee of England, James, Sacrament of USA, Danube (USSR) for phytoplankton studies and Rhine (Germany), Yamuna (India), Nile (Africa), Lower Missouri, Ohio (North America) for zooplankton studies [27].

In contrast to some developed countries, very little attention has been paid to the pollution in Indian rivers. Few scanty references available in literature deal only with ecological point of view in limited zones. Yamuna [28] Cauvery [29], Pawana, Mutha and Mula [30], Jhellum [31], and Khan [32] are some of the rivers in which pollution has been studied in very limited zones.

The dynamics of the river ecosystem for Ganga (India) was determined by Saxena et al. [33] at Kanpur and Ray et al. [28] at Allahabad on the basis of hydrological data collected at these points.

In general, the concentration of coliform bacteria is used as an index of civic pollution. The impact of bacterial population on water quality in various rivers of the world has also been studied. Some pertinent

references are for Hull, Aires and Welsh in England [34-36], Ogilvie, Swift Meduxnepeag [37], Dunbar [38] and Athabasca [39] of Canada, Nile of Egypt [40], Nida of Poland [41], Tisza of Hungary [42], Tigris of Iraq [43] and Sagami of Japan [44].

The studies on bacterial survey of Indian rivers are relatively less. Mathur [45] has investigated the seasonal distribution pattern of coliform bacteria in Yamuna river in Delhi. Bilgrami and Datta Munshi [46] have studied the concentration of coliform bacteria in Ganga river from Patna to Farraka.

Central Pollution Control Board in 1977 initiated a programme on regular water quality monitoring in the Ganga basin under basinwise pollution control task. In the first phase, the water of the river Yamuna was monitored at fifteen sampling stations for the analysis of physico-chemical and bacteriological parameters during the period 1977-1979 [47]. In the second phase, the Central Pollution Board decided to take the cooperation of State Pollution Boards of Uttar Pradesh, Bihar and West Bengal in surveying the water quality of river Ganga with the network of about forty monitoring stations from Hardwar to Diamond Harbour, Calcutta, during early 80's [48]. Nevertheless, this agency did not undertake the mountaneous stretch of the holy river Ganga for monitoring the biological parameters.

The self purification of the river is one of the natural phenomena which controls the microbiol imbalances. Wurhmann [49] summarized the death rate of coliform bacteria in various rivers of the world (Ohio, Missouri, Tennessee and Cumberland) and found that the time required for about 90% decay is close to 10 hrs. The effect of conventional microfauna for the control of bacterial population and rate of the extent of decline of the Escherchia coli bacteria has been studied on marine water by various workers [50-52]. However, the studies to provide precise information on the bacteriological quality in Indian rivers is still lacking in the literature.

Katz [5] has documented some of the effects of municipal waste water effluents on the water quality, fish population and bottom fauna characteristics of a receiving streams and rivers. It was found that the addition of effluents into river influenced the water quality parameters such as dissolved oxygen (low), BOD (high), CO<sub>2</sub>-content (high), fish population (low), bottom fauna (low) and composed primarity of worms (low).

The effect of domestic sewage in the water quality of the Yagyu river in Japan was studied by Ushami et al.[6] and he found that the water pollution is caused by the accumulation of COD, nitrogen, phosphorus,

bacteria and protozoa in a stagnant areas formed by dams in the middle stream. The impact of industrial effluents on the water quality of Welland river, Ontario, which was receiving large quantity of N and C containing inorganic compounds was the subject of interest in 1980's. It was found by Dickman et al. [7], that the areas receiving the waste water discharges were completely devoid of all life except bacteria. Studies on the properties of water in river Isar between Muench and Moosburg, Germany [8] indicated that the discharge of waste water from the Muench sewage plant-II requires immediate attention with respect to complete denitrification and very extensive retention of settleable solid before mixing in the river. Biological studies of fishing in the Upper Weser River in Germany [9] indicated that the effluent of industrial waste water causes sickness in and ultimate death of the fishes. Sickness of the fish is intensified at the chloride concentration of 3000 mg/l. The effect of municipal and industrial waste water in river Bosna (Yugoslavia) in the region from Doboij to Modria was studied by Bendic and Stojanta [10] and it was found that the pollution load was 63.678 pollutant units/inhabitate. An intensive survey on the effect of sewage effluent discharge in San Gabriel and Bushy rivers in Texas, USA indicated that such mixing caused a very low levels of dissolved oxygen below the discharge points. The extensive nutrient



levels stimulated the growth of benthic algae.

Similar studies on various rivers like, Leba [53], Lupawa [54], Slupia [55], Wieprza river basin [56] of Poland, Nagase [57] and Yoda [58] of Japan, River of Southern Missouri, USA [59] and Halifax of Florida, USA [60] have also been carried out. The studies on pollution in the last decade attracted world wide attention and not only the above mentioned rivers but several others important rivers were also studied for the physico-chemical and biological characteristics [11-16,61].

The long term study on the effect of seasonal variations on water quality in the river Rhine has been appeared in the literature recently [62]. The analysis of 240 water samples and of 79 suspended matter samples collected during a 10 year period indicated a complicated pattern of the major dissolved ions due to pollution. A simple steady state model is used to estimate the contribution of aluminosilicate rock and of lime stones to the total denudation. The bacteriological water quality of a multiuse catchment basin on Avalon Peninsula (New Foundland) and in potable water have also been analysed [63,64]. Seasonal fluctuations in population numbers were observed in all samples.

The effect of body size, depth and sampling site on spatial heterogeneity were examined in the zooplankton

community of a small lake by Bernadette et al. [65]. Analyses were performed by regression analysis of 27 sets and results were presented. A carbamate insecticide, 'Carbaryl' was applied in spring to concrete ponds and it was found that at 1.0 ppm, all Zooplankton species were killed [66]. Similarly an organophosphorous insecticide, 'Temphos' was found effective in eliminating almost all zooplankton at a concentration of 500 mg/l [67].

Effect of industrial discharges on the ecology of phytoplankton in the rivers also attracted attention in the last few years. Joy et al. [68] studied the ecology of phytoplankton production in the river Periyar (India), which receives the effluent discharge from a dozen industrial units. Spatial variation of phytoplankton composition, chlorophyll pigments and physico-chemical features of water were observed. The water of river Ganga (India) was also found quite good before entering in Balia city (Uttar Pradesh) while it was found polluted due to the mixing of sewage and industrial effluents at down stream [69].

Shankar et al. [4] made a comparative study on BHEL, Jwalapur drain, Naisota and Lalta Rao sewer drains at Hardwar and found that the colour of sewage was black with pungent smell. The physico-chemical characteristics of water at various selected points were determined. The

Rishikesh sewage was analysed by Chaudhuri [70] and it was found that it carried a high BOD (average 233.1 mg/l) and low dissolved oxygen (0.8 mg/l). The values of COD, MPN, SPC and turbidity were also found to be quite high (average values : 275.5 ppm,  $236.7 \times 10^4$  per ml,  $151.6 \times 10^5$  per ml. and 252.2 NTU respectively).

It is, therefore, clear that pollution in water by mixing the effluents from the industries as well as sewage causes severe pollution problems and requires an immediate attention. Out of the several treatments explored for the treatment of effluents/sewages, use of the hydrophytes has been suggested several decades ago inspite of the fact that it requires several days treatment. Hydrophytes have been found to possess the potential to lower down the pollution load in the polluted water. Mostly these hydrophytes are aquatic weeds and are generally considered as troublesome plants to humans. Fresh-water macrophytes grow naturally in water bodies polluted by nutrients. Many aquatic plants utilize these nutrients and produce large amount of biomass which can be used for some beneficial purposes. The concept of using aquatic plants for treating waste water is gaining attention of local and state agencies these days in various parts of USA (California, Florida, Mississippi, Louisiana and Texas). Aquatic macrophyte system can be effectively used for reducing pollutant

levels in water bodies [1,4,71-73] and the biomass of aquatic macrophytes are used for production of gaseous fuel [74,75] and feed and fibers [76]. Aquatic plants which possess capability of reducing pollution on one hand also suffers from the disadvantages, particularly by Eichhornia crassipes that it grows wildly and the magnitude of the area under its cover can be judged by the remote sensing techniques [77]. Eichhornia crassipes has also been named as Bengal Terror [78], curse of Bengal [78], Blue devil [79] Million Dollor Weed [80] and Cinderella of the plant world [81].

Sioli [82] has observed that water hyacinth roots eliminate some amino acids and carbohydrates in the surrounding water and that the excretion of organic substances may be one of the possible reasons for rich community of aquatic animal life.

Despite several efforts with different control measures and heavy expenditure, the scientists have hardly received any success in controlling the aquatic weeds specially E. crassipes, which provides food to animals and also provides fibre, fuel and fertilizers. Thus, scientist these days are willing to accept aquatic plants specially E. crassipes as their trusted partner in solving the problems of water pollution. Eichhornia crassipes has also been claimed as a future scavenger for various organic and inorganic pollutants from waste

waters. There is an ever growing interest in this plant around the world for reducing pollution in water. Among all the aquatic weeds, E. crassipes is well known for its ability to lose water rapidly through its leaves during transpiration. Due to this property, the plant also derived its name in Hindi as 'Samudra Sokh' (one which is capable of absorbing the ocean). A number of experimental studies have shown that the water loss through evapotranspiration from water hyacinth does not exceed the loss from open water surface. The value reported for the ratio of evapotranspiration (Et)/evaporation (E) and rates  $[Et/E]$  are very divergent. Dass [83] reported the water loss by water hyacinth to be 5.78 to 9.84 (average  $7.76 \pm 1.35$ ) times higher in different seasons as compared to open surface of water bodies.

The growth rate of water hyacinth in India is different at different places. The total annual average net growth of water hyacinth is  $1039/m^2$  and max. rate is  $3.89/m^2/day$  [84]. In Assam, the maximum growth rate was determined as  $9.79/m^2/day$  [83], while in areas near Jaipur, still higher growth rate has been observed in a sewage receiving entrophic water bodies [85,86].

Several experimental studies in laboratories as well as in fields have shown that water hyacinth reduces the quantity of suspended particles, algae, dissolved impurities, nitrogen, phosphorous and other nutrients,

redox potential, BOD, COD, turbidity, faecal coliform bacteria and organic carbon content in water bodies [87-94].

Since the first report published by Pirie in 1960 [95], a large number of reports have appeared in the literature on the possibility of extraction of proteins from water hyacinth and its use for combating protein deficiency in human beings [96]. The value of protein content determined from nitrogen estimations, however, varies widely from 8% of dry weight (0.85% fresh wt) to 5.3 - 5.7% dry weight [97]. Various essential and non-essential amino acids content determined in water hyacinth are summarized in Table 2.1.

**Table 2.1** Essential and Non-essential Amino acids Content (%) in Crude protein of Water hyacinth

Methionine	0.73	Arginine	2.98
Cystine	11.60	Histidine	1.90
Phenylalanine	4.72	Glycine	5.14
Tyrosine	2.98	Alanine	5.59
Threonine	4.32	Serine	5.59
Lysine	5.34	Aspartic Acid	17.37
Isoleucine	4.32	Glutamic acid	9.29
Valine	0.27	Proline	4.33
Leucine	7.20		

A number of workers have shown that aquatic plants also absorb and accumulate a large number of heavy metals (Cu, Cd, Ni, Ag, Cr, Fe, Zn, Mg, Mn, As etc.) and organic pollutants, i.e. phenolic dyes, diphoenamide dyes and photographic pollutants etc [92,94,98-101]. Hence, water hyacinth can be used as sensitive biological indicator for continuous monitoring of trace quantities of toxic heavy metals in an aquatic system [98-100].

Biological methods for purification of water and waste water are generally more energy efficient and cost effective than physico-chemical methods. Vascular aquatic plants employing solar energy as a principal energy source have shown the capability of absorption, translocation and/or metabolic breakdown of heavy molecules present in water and waste water [102].

Saksit [102] conducted the field experimental study, batch screening study and continuous study to determine the trace heavy metal ions removal capacity of various plants in effluents and sewage waste. The objective of the field study was to observe the trace contaminant accumulation under natural condition in an effort to determine plant species for screening study. Plant species efficiency potential as determined by preliminary literature evaluation and experience. The plants were divided into three categories, viz.,

floating, submerged and rooted once as shown in table 2.2.

**Table 2.2** Details of aquatic plant species selected by Saksit [102]

Sl.No.	Common Name	Scientific Name	Classification
1.	Bulrush	<u>Seirpus</u> Sp.	Rooted
2.	Rush	<u>Juncus</u> Sp.	Rooted
3.	Arrohead	<u>Sagittaria</u> <u>graminea</u>	Rooted
4.	Water hyacinth	<u>Eichhornia</u> <u>crassipes</u>	Floating
5.	Duck Weed	<u>Lemna</u> <u>minor</u>	Floating
6.	Water bonnet	<u>Pistia</u> <u>stratiotes</u>	Floating
7.	Elodia	<u>Elodia</u> <u>canadensis</u>	Submerged
8.	Contail	<u>Ceratophyllum</u> <u>demorsum</u>	Submerged
9.	Alligator weed	<u>Alternanthera</u> <u>philexeroides</u>	Submerged

The batch screening study was conducted to determine the removal capacity of heavy metal ions by various plant species. Bulrush [Seirpus sp.] was observed as the efficient rooted system. This plant exhibited a very high capacity of removing cadmium (Cd) from aquatic system. Water hyacinth appeared to be the most effective floating species, for trace contaminant



reduction and it showed the highest removal capacity of cadmium. Among the submerged plants Ceratophyllum demorsum showed the highest capability for reducing cadmium.

Plants play more dominant role than aquatic animals in the aquatic system because they have great effect on the aquatic environment and are more adaptable to either harsh or fluctuating environmental conditions. Floating plants have their photosynthetic part at or just above the water surface with roots extending down into the water column. This extensive root system acts as an excellent medium for filtration/absorption and bacterial support potential. Root development has been shown as the function of nutrients available in the water and the nutrient demand of plants [103].

Aquatic system treatment occurs at a comparatively slow rate in essentially unmanaged natural environment, while the aquatic plants, viz., E. crassipes in particular remove the trace heavy metal ions rapidly from an aquatic system. Heavy metal ions are removed from the water and waste water [103] during aquatic treatment by plant uptake (plant adsorption), chemical precipitation (formation of precipitates with insoluble compounds), ion exchange with settled clay and organic compounds and adsorption (on substrate and plant surface). Metal uptake by plants depend upon several

factors including pH, concentration of metal ions, depth of root system of plant and growth rate of plant in the aquatic system. Higher initial concentration of the trace heavy metal ions, extensive root system and growth rate of water hyacinth plant increases the total uptake capacity of the plant.

The water hyacinth plants completely saturated with heavy metal ions are periodically removed and disposed of in the adjacent pit. The useful products like protein, biogas, compost, fiber etc. can also be obtained from the disposed hyacinth [73-76].

Biomass is defined as a quantitative estimate of the total mass of living organisms within a given area or volume. It may include the mass of a population or of a community but provides no information on community structure. As water hyacinth plants are living organisms of plant kingdom, they exhibit characteristics like very fast growth rate, unlimited annual productivity owing to hydrous factors, considerable power of absorbing matters (dissolved or suspended) in liquid media and as a result it has tremendous capability for water purification.

Haller et al. [104] using radioisotope of phosphorus ( $P^{52}$ ), reported that larger amounts of phosphorus were absorbed by water hyacinth from

solutions containing phosphorus. Further, upto 50% of the absorbed phosphorus backed out from the stem and root tissues within six days.

The use of water hyacinth for nutrients removal was first demonstrated in 1967 by Sheffield [91]. He reported 80% reduction in  $\text{NH}_3\text{-N}$  when the aerated effluent passed through water hyacinth pond with retention time of 10 days. Reduction in the nitrate nitrogen occurred upto 81% but only after anaerobiosis was established in 42 days. Phosphates were reduced initially upto 51% but decreased to only 20% after one month due to the release of phosphorus from decaying plants. Clock [105] observed that in a pond with five days detention time, 39-94% of total nitrogen and 61% phosphates were removed by water hyacinth and still greater amounts were removed during the period of vigorous growth.

Miner et al. [106] reported removal of 23 lb/acre ammonia nitrogen and 17 lb/ac phosphate from a sewage lagoon. Bagnall et al. [107] found that water hyacinth removed only 9% nitrogen and 8% phosphorous from secondary treated effluent from a pond with a detention time of 10 hrs. However, the percentage removal increased to 60% with five days detention period. They also observed that crowding, lower temperature and greater water depth resulted in low removal of

nutrients. Further, they recommended regular harvesting of plants at the rate of 25 to 35% per week. Similar observations on depth of water and retention time were made by Cornwell et al. [108] and Minshall et al. [109].

Other factors affecting the removal of nitrogen and phosphorus include release of nitrogen as gas, concentration of nutrients due to evaporation, formation of complex precipitates involving phosphates and growth rate of plants, temperature etc. It has been pointed out that in temperate region, the plants would have to be introduced every year and large populations would be required to remove nutrients within a short time [110].

The capability of pollutant removal by water hyacinth have been found during different studies. The data obtained for the removal of some of the pollutants is quite impressive, viz., 100 mg phenol in 72 hrs. by 2.72 gms dry wt. of water hyacinth [101], 75 mg fluoride [89] and also 80.8% cadmium in seven days [111]. The water hyacinth have also been shown to accumulate abnormally high quantities of cobalt and Iron [111]. In another study, hyacinth plants have also claimed to absorb sulphates and sodium from textile factory wastes [112]. Based on these studies, the use of water hyacinth in sewage treatment has been suggested by many workers [98,99,113-114] particularly for advanced water treatment [115].

Shankar et al. [4] found that Eichhornia plant grown by using effluents of Indian Drugs and Pharmaceuticals Ltd. Rishikesh, Uttar Pradesh, India, were healthy and the number of leaves increased four times within 30 days and was better than that of plant grown using tap water. Besides, there was increase in dissolved oxygen level and decrease in BOD value of effluent. Hardness, alkalinity, turbidity and Chloride on the other hand showed a declining trend.

The use of Chlorella pyrenoidesa and Scenedesmus species in the removal of nutrients from the waste water has also been reported by Tam and Wong [116]. It was found that these plants were capable of reducing more than 80% of total phosphorus and inorganic nitrogen present in settled sewage. Jain et al. [117] used recently Azolla pinnata and Lemna minor for the removal of lead and zinc from polluted water.

Thus, it is clear that plant species have long been used for removal of toxic metal ions from the effluents and waste water and thus are the cheap source for pollution reduction. In the present studies, attempts have been made to use Eichhornia crassipes, Lemna minor and Ceratophyllum demorsum for the removal of pollution load. The physico-chemical characteristics of waste water have also been determined before and after the

treatment. It was observed that species of Eichhornia is particularly useful for reducing the pollution load of sewage and effluent under investigation.





CHAPTER - III  
MATERIALS AND METHODS

### 3.1 DETAILS OF RIVER DHAMOLA

Dhamola river is an important river of Saharanpur city of Western Uttar Pradesh. It originates from Dehradun hilly region (basin of Himalaya) as a natural water source. It flows through the Saharanpur city from North to South. The quality of water of this river, about 5 to 6 Kms before entering into Saharanpur city, is very good and is used for drinking, bathing and agricultural purposes by animals as well as human beings. The altitude of the river Dhamola near Saharanpur is about meters. The flow of this river ranges from 2-3 km/hr during entire year. The length of this river in Saharanpur city is about 12 kilometers and breadth almost 10-30 meters which depends on the mode of flow. The total flow of water is about 500-800 million liter per day. The average depth of the river is about 0.9-1.5 meters. But the depth and breadth of this river is different at different sampling stations. The water of the river Dhamola just before entering the North of Saharanpur city (upstream) is colourless and visually looks pure, whereas, after about 12 kilometers of entering the Saharanpur (down stream), the grey colour of the water can be easily seen. Essential micro and



macro nutrients of the water in the river are very much influenced by the mixing of waste discharge. Hence, the contamination by sewage and other waste products causes pollution by disturbing the water quality of river Dhamola. Since, Saharanpur is an important city, it was considered interesting to explore the cause of pollution in Dhamola river so that the role of industrial effluents and sewage on pollution can be evaluated.

The quality of Dhamola river water before entering the Saharanpur city is very good and the flora and fauna (aquatic) survive magnificiently. But, after the mixing of industrial effluents and sewage water, the aquatic life was found to be disturbed. The origin of river Dhamola and map of Saharanpur city, position of industries and points of mixing effluents and sewage discharge into the river Dhamola are presented in figures 3.1 and 3.2.

### 3.2 SAMPLING STATIONS

Sampling stations were selected after an extensive survey of the area on the basis:

- that the site represents the state of aquatic system including inputs, withdrawals and other related features.
- that it should be accessible in all seasons.

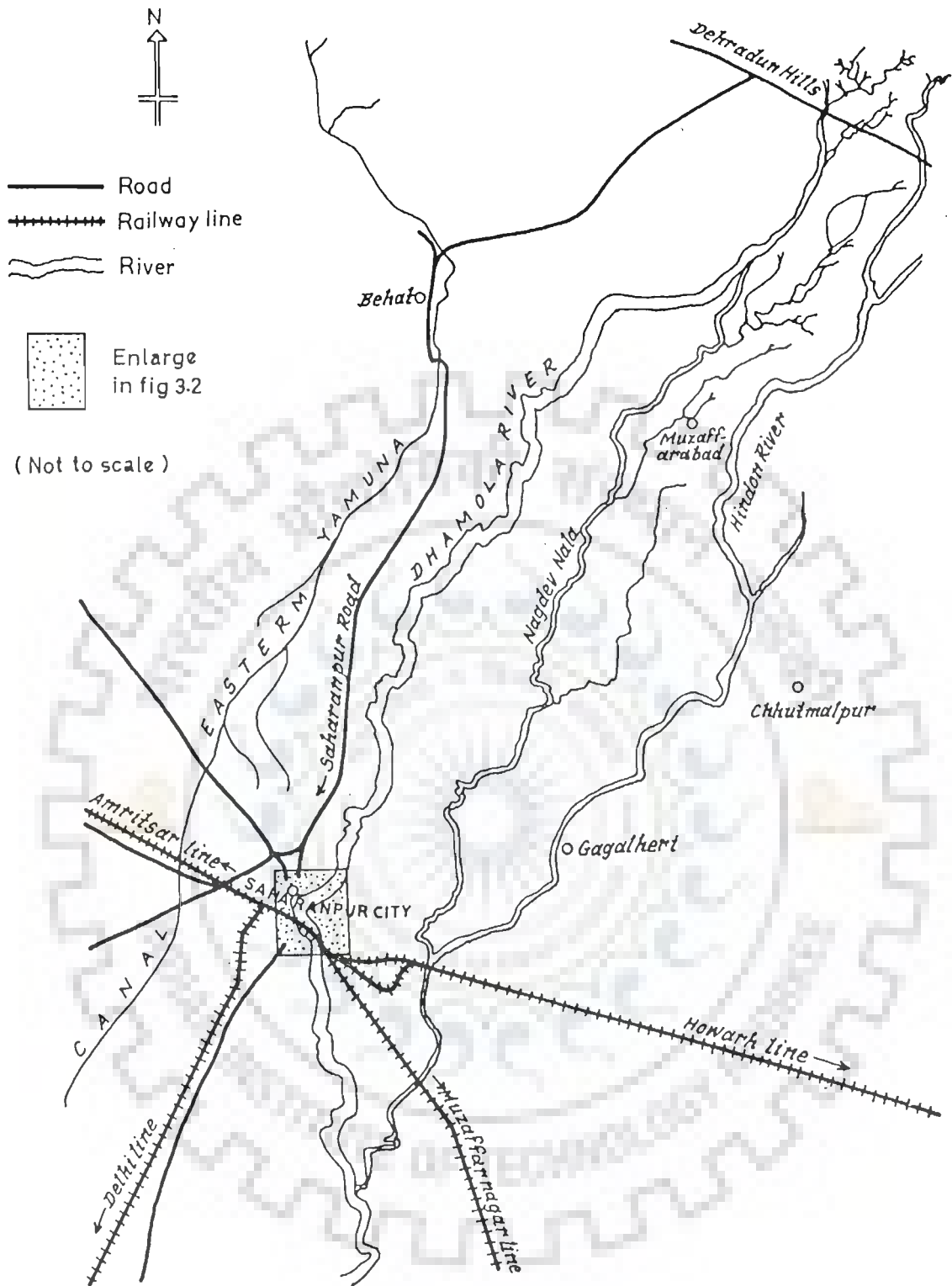


Fig. 3.1 Area map for the origin of River Dhamola

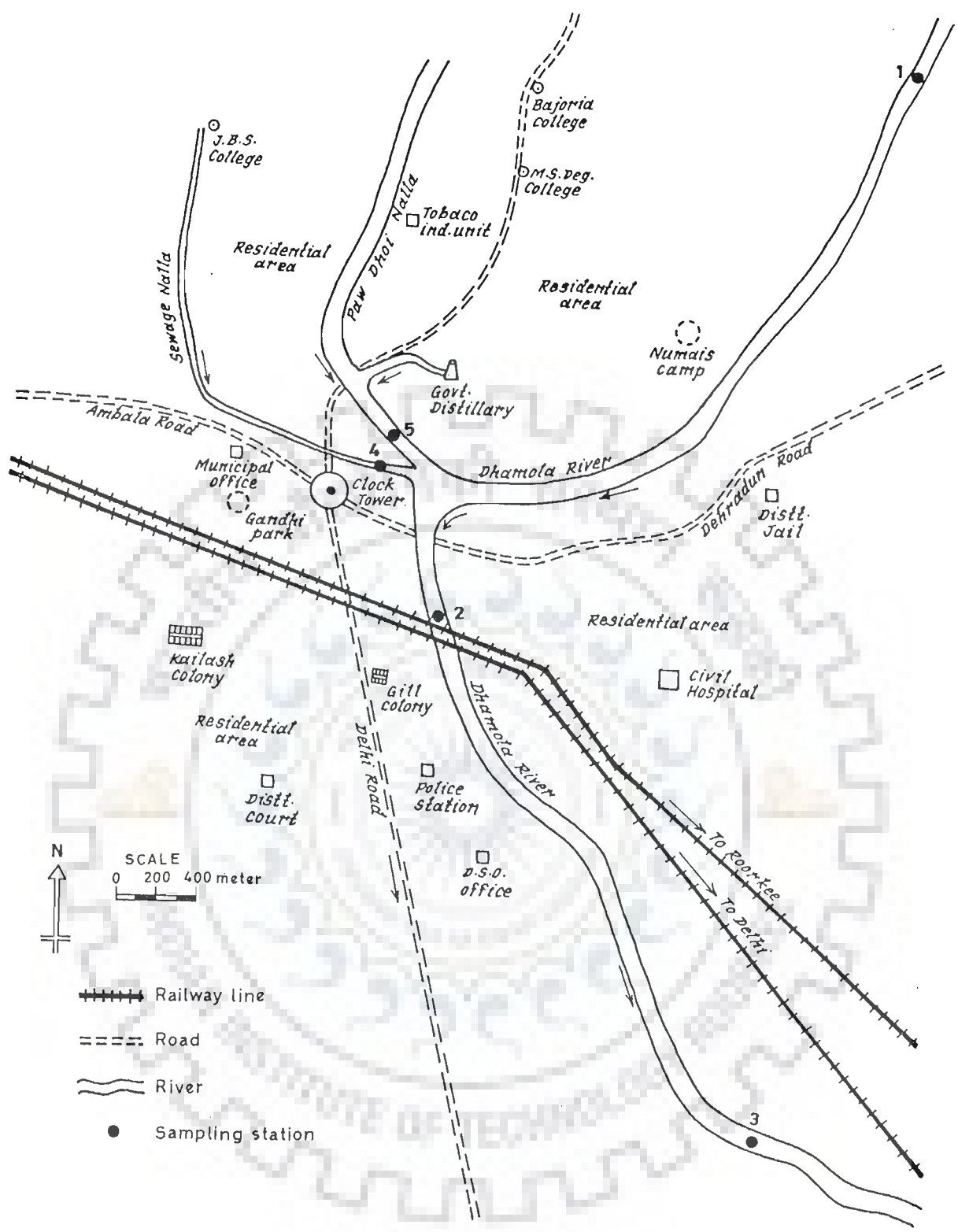


Fig. 3.2 Area map of river Dhamola in Saharanpur city

- that it should have uniformity of concentrations of pollutants across the river section implying well mixed conditions, and
- that it should be concerned with the population of that place.

Five sampling stations covering the stretch of about 12 Kms of Dhamola river at different areas of Saharanpur were selected as shown in Fig. 3.2. Points 1,2 and 3 provided water of the river Dhamola whereas 4 and 5 are the sampling stations of sewage water and industrial effluent respectively. The details of the selected sampling station are as follows:-

### 3.2.1 Sampling Station - 1

The first sampling station was selected before the entering of Dhamola in Saharanpur city. This sampling station is represented by no.1 in fig. 3.2. At this point the depth of the river is about 1.2 meters and breadth about 20-25 meters. The flow of water at this point is about 2 kms/hr. The water looks visually pure. A photograph of this sampling station is presented in plate - 1.

### 3.2.2 Sampling Station - 2

The second sampling station was selected after 2 km of the mixing of effluent waste discharge (industrial as

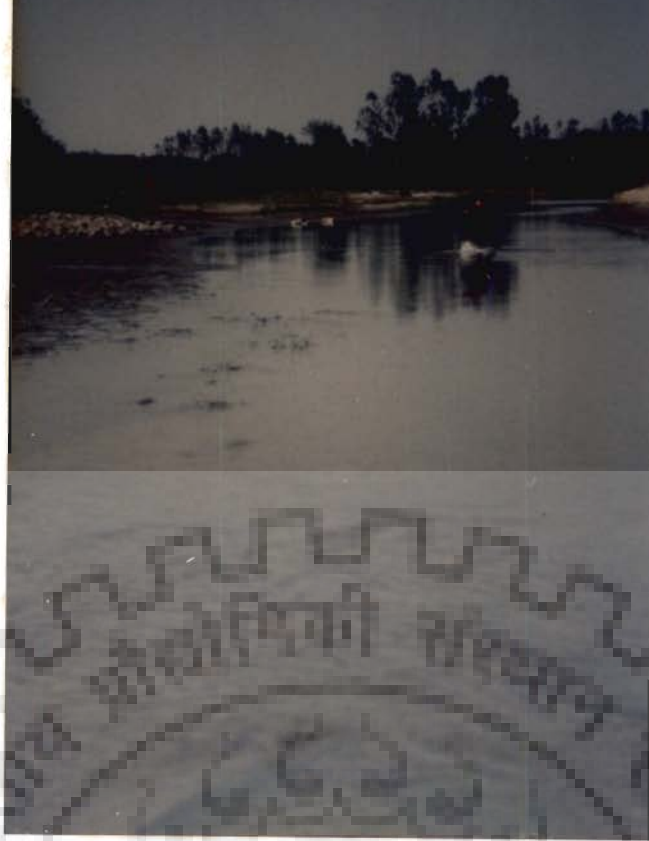


Plate 1 Dhamola river before entering into Saharanpur City (sampling Station 1)



Plate 2 Mixing of waste water into the river Dhamola

well as domestic) at about 7 kilometers from sampling station 1. The water at this point is dark grey in colour with pungent smell. The depth of the river at this point is about 1.35-1.5 meters and breadth 20-30 meters. The usual flow at this point is also about 2-3 km/hr. The altitude at point - 2 is 266.67 meters. Photographs of Dhamola at sampling point - 2 are presented in plates 2 and 3.

### 3.2.3 Sampling Station - 3

The third sampling station further down stream was selected outside the city about 5 kms apart from sampling point - 2. The depth and breadth at this sampling station is practically similar to point - 2. The flow rate of water was also found to be same as that of points - 1 and 2. The elevation of this point is 265.37 meters. Photographs showing details of this sampling station are presented in plates 4 and 5.

### 3.2.4 Sampling Stations 4 and 5

The fourth and fifth sampling stations selected were the places where the sewage water and the effluent of the industries respectively as shown in Fig. 3.2. Though the waste water of industry and sewage merge in Dhamola at the same point, however, the samples of both were collected separately about half a kilometers before the mixing in Dhamola. The exact discharge from the industries and sewage could not be determined. However,

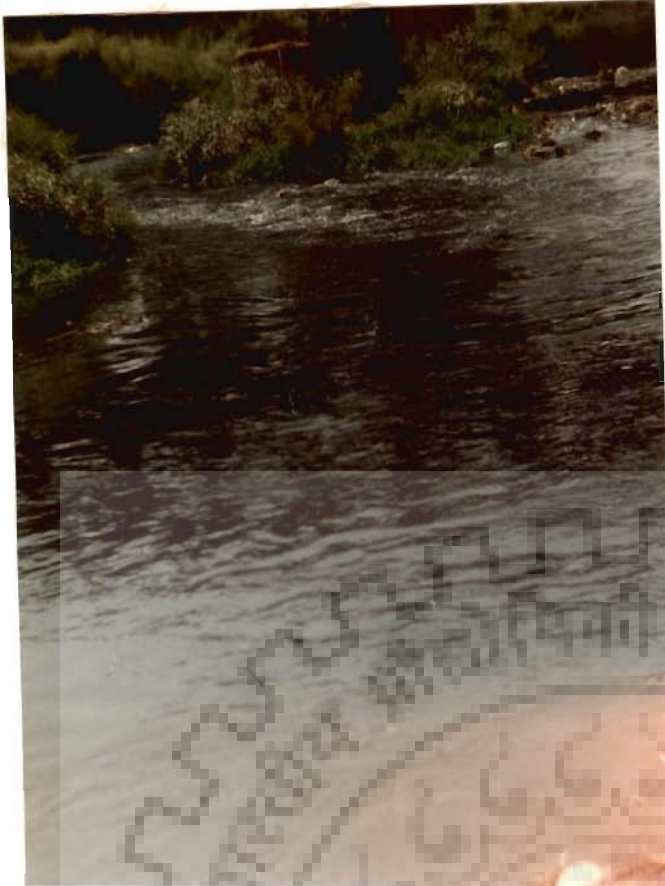


Plate 3 Sampling station 2

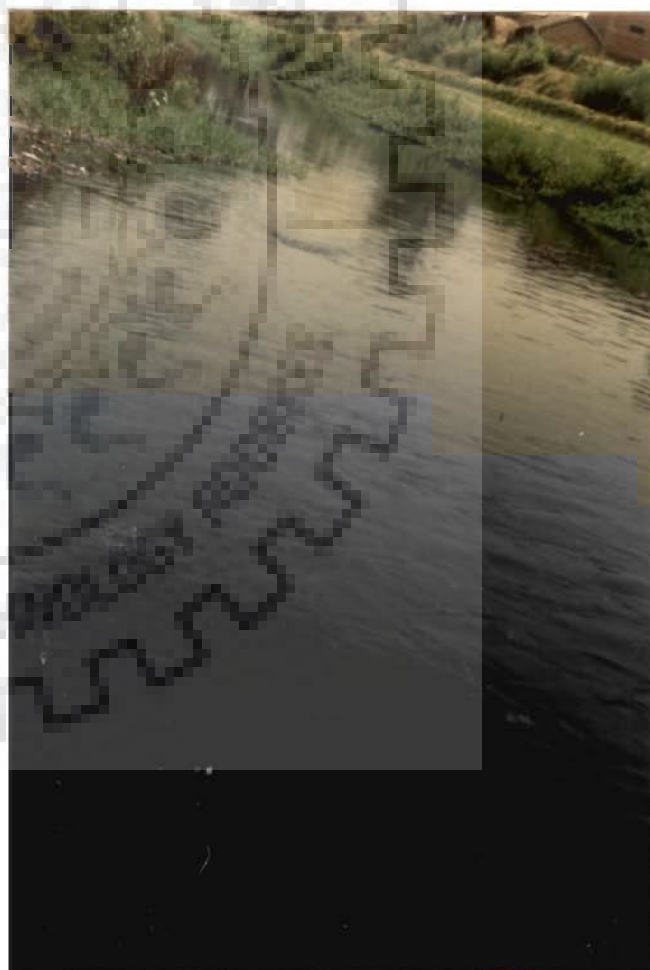


Plate 4 Sampling station 3, seven kms downstream of the waste water mixing



Plate 5 Sampling station 3 (close up)



Plate 6 Domestic sewage (sampling station 4)



the approximate values mentioned by the related authorities were 80-100 million liters per day for sewage waste and 50-60 million liters per day for industrial effluent. Some selected photographs taken at various points are presented in plates 6 and 7.

### 3.3 Sampling Program

Samples from points 1 to 5 were obtained on 15th/16th of every month from July, 1988 to June, 1989. From July 1989 onwards, the frequency was reduced to once in three months.

### 3.4 Sample Collection

For physico-chemical analysis, the samples were collected in large polyethylene containers. These containers were cleaned with dilute NaOH before sampling and again washed several times with water for sample collections. Samples were collected at the depth of approximately 0.5 meter below the surface of running water. Handling of samples was done very carefully to minimize the occurrence of any change during storage and transportation. Samples for the determination of dissolved oxygen were collected in 300 ml BOD bottles with the help of a sampler. The bottles were filled completely upto the brim. These samples were preserved by adding 1.0 ml each of alkaline potassium iodide solution and mangnous sulphate solution at the sampling sites.



Plate 7 Industrial effluent (sampling station 5)



As the water at the top of the fast flowing river has low plankton density, the samples for plankton studies were concentrated in order to get count in an appropriate range. 100 litres of water collected by a bucket of 10 litres, was filtered through a plankton net (bolting silk No. 25, mesh size 55 micron) to collect the plankton samples. The total community obtained was transferred to a clean 250 ml glass bottle and the final volume was made upto 100 ml with the filtered water. Generally the preservation was avoided as it may result in the transformation of shape and size of some of the individual organisms, making the identification difficult. Wherever it was necessary, two parallel samples were taken, one without preservative and the other with preservative (4.0 percent neutralized formalin).

Water samples for bacteriological examination were usually taken from a depth of 0.3-0.5 meter below the surface. Bacteriological water samples were collected in sterile narrow mouthed glass bottles of 250 ml capacity with ground stoppers. The details of various containers used for collection of samples, their optimum storage time and the method for the preservation related to various water characteristics are summarised in Table 3.1.

Table 3.1 Summary of special sampling or handling used

Parameters	Container	Optimum storage time	Method of Preservation
Turbidity	P	With in a week	Refrigeration
Total solids	P	With in a week	Refrigeration
pH	P,G	Analysed at site	-
Total hardness	P	Analysed same day	Refrigeration
Chloride	P	Analysed with in a week	Refrigeration
Dissolved oxygen	G	Fixed at sampling site, analysed same day	Refrigeration
BOD	P,G	Analysed and incubated same day	Refrigeration incubated in BOD incubator
COD	P.G.	Analysed same day	Acidified with concentrated $H_2SO_4$
MPN (Coliform bacteria)	G	Analysed same day	Refrigeration
SPC(Coliform bacteria)	G	Analysed same day	Refrigeration
Algae	G	Analysed same day	Refrigeration
Diatoms	G	Analysed same day	Refrigeration
Protozoan	G	Analysed same day	Refrigeration

P = polyethylene container, G = glass bottle.

## 3.5 SIGNIFICANCE OF PHYSICO-CHEMICAL PARAMETERS

### 3.5.1 pH

pH is an important parameter of both natural and waste water as it is a measure of acidity or alkalinity. The pH range for the existence of most biological life is quite narrow. The industrial waste, due to the presence of acids and alkalies in it, alter the pH of the water at mixing point. It is thus very important that the excessively acidic or alkaline waste should not be discharged without neutralizing the acidity or alkalinity of the waste water. The determination of pH value also helps in the chemical and biological treatment of the waste water.

### 3.5.2 Turbidity

Turbidity is an expression of the optical property that causes light to be scattered and absorbed rather than transmitted in straight lines through the sample. In water, turbidity is caused due to the presence of suspended matter like, silt, clay, finesand, inorganic and organic matter, soluble coloured organic compounds, plankton and other micro organisms. It is useful in describing the clarity of water. The turbidity is measured in Nephlo Turbid Unit or Jackson Turbid Unit.

### 3.5.3 Total Solids

Total solids value for a water sample is defined

as the amount of residue left after the evaporation of a known volume of sample and its subsequent drying in an oven at a fixed known temperature. The total solids determined in a water sample include the dissolved material, e.g., soluble organic/inorganic compounds, and suspended matter (plankton, clay, and other micro and macro organisms). Water with high total solids generally affect water quality adversely as it has inferior palatability. A limit of 500 mg dissolved solids/litre is generally acceptable for drinking water.

#### 3.5.4 Dissolved Oxygen

Dissolved oxygen is an important test for determining pollution in water. The level of dissolved oxygen in natural and waste water depends upon the physico-chemical and biochemical activities in the water body. It directly affects the aquatic life in water because it is necessary for all the aerobic micro as well as macro organisms. It is measured in mg/l. and calculated using the expression :

$$\text{DO (mg/l)} = \frac{\text{volume in ml of } 0.025 \text{ N Na}_2\text{S}_2\text{O}_3 \text{ used in titration}}{1000}$$

#### 3.5.5 Biochemical Oxygen Demand

BOD is the amount of oxygen required by micro-organisms for the degradation of organic matter under aerobic conditions. The waste water usually does

not contain sufficient amount of oxygen to maintain aerobic condition during degradation. For complete stabilization of organic matter, required amount of oxygen is taken as a measure of its organic content. It is measured in mg/l unit. The relation used for calculation of BOD is :

$$\text{BOD (mg/l)} = \frac{[D_1 - D_2] - [B_1 - B_2] f}{P}$$

where

$D_1$  = DO of diluted sample immediately after preparation (mg/l).

$D_2$  = DO of diluted sample after 5 days incubation at 20°C (mg/l).

$B_1$  = DO of seed control before incubation (mg/l).

$B_2$  = DO of seed control after incubation (mg/l).

$P$  = decimal volumetric fraction of sample used.

$f$  = ratio of seed in diluted sample to seed in seed control (% seed in diluted sample) / (% seed in seed control).

### 3.5.6 Chemical Oxygen Demand

Chemical oxygen demand is the required amount of oxygen in water for the complete oxidation of organic matter. It is directly related to the chemical changes in natural water bodies due to the mixing of certain waste products (industrial as well as domestic). It is measured in mg/l units. It is determined using the expression :

$$\text{COD (mg/l)} = \frac{[a - b] N \times 8000}{\text{ml of sample}}$$

where

a = ml of  $\text{FeSO}_4(\text{NH}_4)_2\text{SO}_4$  for blank.

b = ml of  $\text{FeSO}_4(\text{NH}_4)_2\text{SO}_4$  for sample.

N = Normality of  $\text{FeSO}_4(\text{NH}_4)_2\text{SO}_4$  solution.

### 3.5.7 Carbon Dioxide (Free)

Pure water usually contains less than 10 mg free carbon dioxide per litre. The  $\text{CO}_2$  content of water is an important parameter as it contributes significantly to corrosion. Free carbon dioxide occurs in water due to the combustion of carbonic matter present in water bodies. It can be determined by potentiometric titration or with phenolphthalein indicator. In the present studies phenolphthalein indicator method was employed. The calculation used is as follows :

$$\text{CO}_2 \text{ (mg/l)} = \frac{A \times N \times 44000}{\text{ml of sample}}$$

where

A = ml used of the titrant

N = Normality of titrant

### 3.5.8 Chloride

Chloride is another important water quality



parameter which gives an idea of the hardness and taste of water. It is present in the form of chloride ( $\text{Cl}^-$ ) ions which are the major inorganic anions in water and waste water. The amount of chloride present in water varies from 10 mg/l to 500 mg/l in natural water bodies depending upon the source. The high value of chloride in water indicates pollution. The municipal sewage is the primary cause of increase in chloride concentration as NaCl is a common constituent of diet and passes unchanged through the digestive system. It is represented as mg/l and can be calculated using the formula :

$$\text{Cl}^- \text{ (mg/l)} = \frac{(\text{ml of AgNO}_3 \text{ used for sample} - \text{ml of AgNO}_3 \text{ used for blank}) \times 1000}{\text{ml of sample}}$$

### 3.5.9 Total Hardness

Hardness in the water is due to the presence of salt of calcium and magnesium. Hence, the total hardness is defined as the sum of calcium and magnesium concentrations. It poses problem of scale formation in boilers. The hardness of water may vary from zero to about few hundred mg/l depending on the source of water. It is expressed as calcium carbonate in mg/l. It is determined by volumetric method using EDTA. The hardness is calculated using expression :

$$\text{Hardness (mg CaCO}_3\text{/l)} = \frac{A \times B \times 1000}{\text{ml of sample}}$$

where

A = ml of titrant for sample

B = mg CaCO<sub>3</sub> equivalent to 1.0 ml EDTA.

### 3.6 METHODS USED

A brief summary of the techniques used and its principle for physico-chemical and biological parameters is presented in Table 3.2.

### 3.7 CHEMICALS AND APPARATUS USED

The chemicals used for determining physico-chemical and biological parameters of water were obtained from the following sources : E.Merck (India) Limited, Bombay, (Ammonia buffer, Potassium chromate indicator, Silver nitrate, Manganese sulphate, Potassium iodide, Sodium thiosulphate, Starch, Phosphate buffer, Magnesium sulphate, Ferroin indicator, phenolphthalein, Potassium hydrogen phthalate and Sodium hydroxide), HiMedia, Bombay (Lactose, Peptone, Sodium chloride, Eriochrome black T-indicator) and BDH chemical Division, Bombay) (Sulphuric acid, Ferric chloride, Calcium chloride, Potassium dichromate and Ferrous ammonium sulphate).

Table - 3.2 Summary of Analytical Methods

Parameters	Technique	Principle/Apparatus
Turbidity	Nephelometry	Nephelometry [118]
Total solids	Gravimetric	Evaporation at 103°C - 105°C
CO <sub>2</sub> (free)	Volumetric	Phenolphthalein titration, titrant 0.1 N NaOH [118]
Hardness	Volumetric	Complexometric titration [118] titrant 0.01 EDTA, indicator-Erichrome black-T indicator, using ammonia buffer.
Chlorides	Volumetric	Argentometric titration[119] titrant M/35.5 AgNO <sub>3</sub> , indicator potassium chromate (K <sub>2</sub> CrO <sub>4</sub> ).
Dissolved Oxygen	Volumetric	Winkler's Method [119], titrant N/80 Sodiums thiosulphate indicator starch (fresh).
BOD	Volumetric	Winkler's method [118] 5 day incubation at 20°C.
COD	Volumetric	Reflux method followed by redox titration, titrant -0.1 N Ferrous ammonium sulphate [FeSO <sub>4</sub> (NH <sub>4</sub> ) <sub>2</sub> SO <sub>4</sub> ], indicator Ferroin [118].
MPN	McConkey	Three tubes of McConkey broth (purple) incubated for 24 hrs and 48 hrs.
SPC	dilution	Standard agar plate (1.5%) incubated for 24 hrs [118].
Phytoplankton	Microscopic	Sedgwick-Rafter Counting Cell, 1/100 dilution [118].
Zooplankton	Microscopic	Sedgwick-Rafter Counting Cell, 1/100 dilution [118].

The glass apparatus used for physico-chemical and biological parameters were sterilized in an autoclave (SEW India, Delhi). The pH was determined using Global, DPH-500 pH meter. HACH Nephelometer (turbidity meter model - 2100 A) was employed for determination of turbidity. The oven (model-USW 143), BOD incubator (Narang Scientific Works, New Delhi), Sedgwick-Rafter Counting Cell, Plankton net (No.25), S-J Sampler, colony counter and microscope (model - 47, Beck, London were used.

### 3.8 BIOLOGICAL PARAMETERS

#### 3.8.1 Plankton

The counting of different phytoplankton and zooplankton, e.g. algae, diatoms and protozoan was achieved from the 100 ml sample in which total community was suspended after filtering the water through plankton net. The device employed for counting was based on the Sedgwick Rafter (S-R) counting cell at area 50 x 20 x 1 mm with a total capacity of 1.0 ml sample. The sample was mixed thoroughly using a shaker at the speed of 25 r.p.m. 1.0 ml of the sample was then transferred by using a large bore pipette to Sedgwick Rafter cell cavity. All the plankton population (algae, diatoms and protozoan) were counted in 100 random fields (chambers) simultaneously under 40 times magnification with a Metzger compact inverted microscope [120]. The average



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number of plankton count per chamber was multiplied by the total countable factor to obtain the total number of individuals per species in the concentrated samples. The counting of one sample was made atleast three times to get a reliable information. For the identification of plankton species, high magnification (450 times) was used. The planktons were identified with the help of following Edmondson [121], Smith [122], Desikachary [123] for phytoplankton and Pennak [124], Calaway and Lackey [125] and Needham and Needham [126] for zooplankton.

### **3.8.2 Bacterial Population**

The water samples collected in 100 ml glass bottles were immediately tested for coliform and total heterotrophic bacterial, e.g. for the incubation in the field, a portable, battery operated incubator was used. The temperature of incubation was  $35 \pm 0.5^{\circ}\text{C}$  and the time was 24 hrs. In some cases a multiple of 24 hrs as required for the different test was used.

#### **3.8.2.1 Coliform Bacteria (MPN)**

Population of total coliform was enumerated by MPN techniques as per Standard Methods [118]. The tubes of  $M_c$ Conkey broth (purple) were used for each dilution. The unit to determine the MPN count is number of bacteria per 100 ml.

### 3.8.2.2 Standard Plate Count of Heterotrophic Coliform Bacteria

The total heterotrophic coliform bacteria were enumerated in terms of standard plate count (SPC) bacteria using nutrient agar (1.5%) as culture media. The water samples collected were diluted in saline (0.9% NaCl) and checked on agar-agar plates for the bacterial colony measurement by the inoculation method. If the number of colonies were more than 300, the samples were further diluted and left for 24 hrs and the colonies were counted again. The counting of SPC is generally number of bacteria present per ml.

## 3.9 STATISTICAL ANALYSIS

### 3.9.1 Correlation Coefficient

For the purpose of this investigation, we have utilized a comprehensive data observed for the three distinct sampling sites at river Dhamola during the period of two years from July 1988 to June 1990. In order to calculate the correlation coefficients among the water quality parameters of river Dhamola, we have used all the fourteen water quality parameters determined in the present studies. These are turbidity total solids, pH, hardness, chloride,  $\text{CO}_2$ , DO, BOD, COD, MPN, SPC, algae, diatoms and protozoans.

Let  $x$  and  $y$  are two variables (water quality parameters in the present case) and  $x_1, y_1$  be  $n$  pairs of the observed values of these variables ( $i = 1, 2, 3, 4, \dots, \dots, n$  and  $n = 48$  in the present case). Then the correlation coefficient 'r' between the variables  $x$  and  $y$  can be represented by the relation :

$$r = \frac{\sum YX}{[\sum X^2 \cdot \sum Y^2]^{\frac{1}{2}}} \quad \dots\dots 1$$

where

$$X = (x - \bar{x}), \quad Y = (y - \bar{y}) \quad \dots\dots 2$$

$$\text{and } \bar{x} = [\sum x]/n, \quad \bar{y} = [\sum y]/n \quad \dots\dots 3$$

and all the summation [  $x$  or  $y$  ] are to be taken for  $n$  number of observations.

### 3.9.2 Parameters of Straight Line

If the correlation coefficient 'r' between two variables  $x$  and  $y$  is fairly large, it can be calculated that a linear relationship of the form:

$$y = A + Bx \quad \dots\dots 4$$

exists between the two variables. In this equation  $A$  and  $B$  are the parameters of the straight line and the value of  $A$  and  $B$  are constant for a particular line. It has been showed by Wonnacott and Wonnacott [127] and Spiegel [128] that the best estimates of the parameters  $A$  and  $B$

can be represented by the equation :

$$B = \frac{\sum (x-\bar{x}) (y-\bar{y})}{\sum (x-\bar{x})^2}$$

i.e.

$$B = \frac{\sum x.y - \bar{x}.\sum y}{\sum x^2 - \bar{x}.\sum x} \quad \dots\dots 5$$

$$A = \bar{y} - B\bar{x} \quad \dots\dots 6$$

where

A is the intercept and

B is regression coefficient i.e. slope of the line.

The summations are to be taken from 1 to n as in the earlier case.

### 3.10 BIOMASS OF HYDROPHYTES

Total biomass of hydrophytes was determined in the terms of dry mass of plants. A known quantity of plant was taken and dried in an oven at 110°C for 15 minutes. It's mass was then determined. The temperature of the oven was then reduced to 50-55°C and the plants were dried for a week. The mass of dried plant was then determined. Biomass was then calculated as the percentage of the dried weight.

$$\text{Change in Biomass} = \text{Final biomass} - \text{initial biomass}$$



Final biomass = Mass obtained after one week

Initial biomass = Mass obtained after 15 minutes

### 3.11 REMOVAL OF POLLUTION LOAD

The waste water samples collected for domestic sewage and industrial effluent (collection points 4 and 5, Fig. 3.2) were transferred into 60 litres polyethylene container and allowed to settle for 2-3 hours. After the sedimentation, the waste water was poured into the aspirator bottle (20 litre) and allowed to flow until all the jars shown in Fig. 3.3 were atleast half filled. The stopper of all the jars was then closed so that the flow of sample from one jars to another could be stopped. The following hydrophytes were taken in the jars:

- Stage I - Eichhornia crassipes
- Stage II - Lemna minor
- Stage III - Ceratophyllum demorsum
- Stage IV - Eichhornia crassipes and Lemna minor
- Stage V - Lemna minor and Ceratophyllum demorsum

A sketch of the plant species used in stage I to III is presented in fig. 3.4. Three such arrangements were made so that the data can be obtained in triplicate.

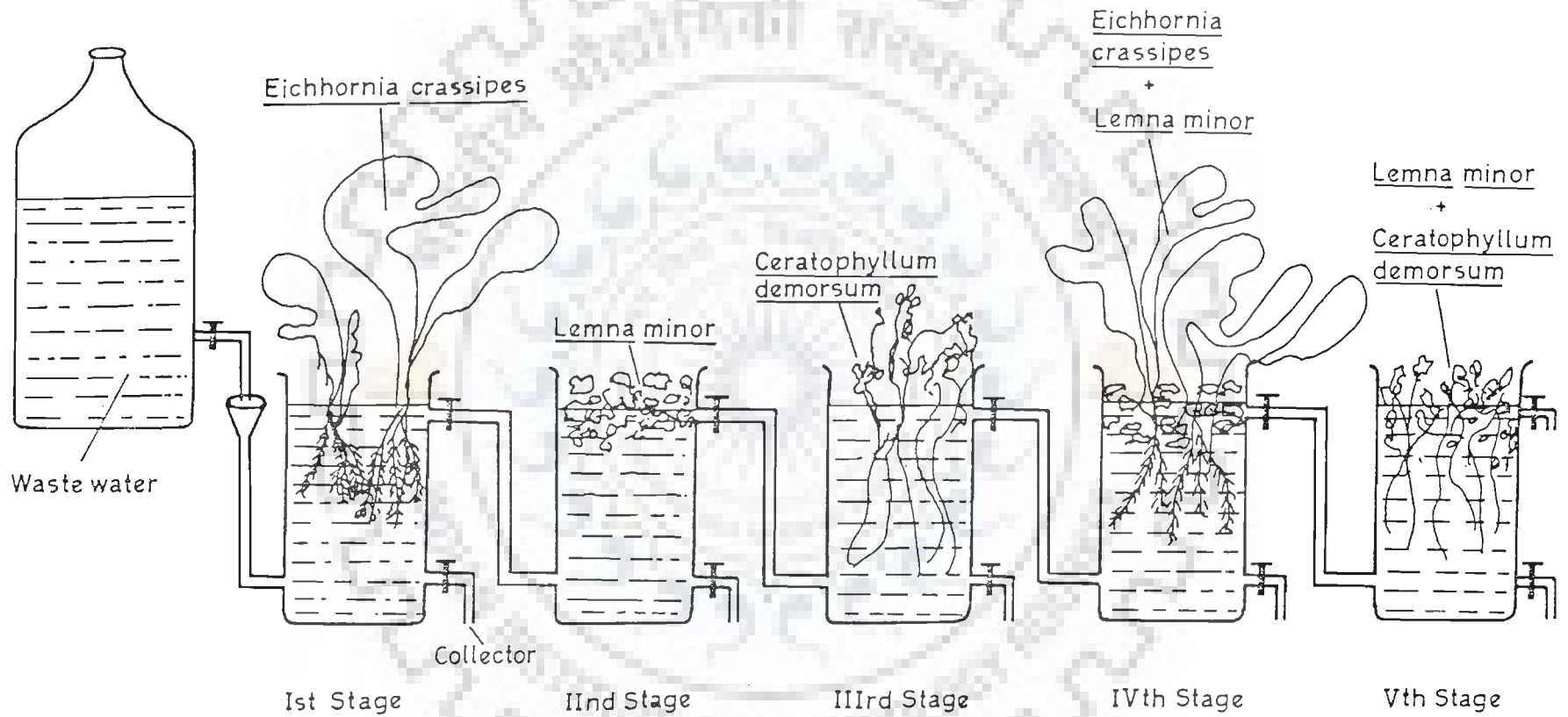


Fig. 3.3 Details of the treatment process used for the water samples with the selected hydrophytes



Fig. 3.4 Outlines of the different species of hydrophytes used

The hydrophytes added at each stage was fixed in such a manner so that the surface area of the sample was totally covered. All the experiments were performed at room temperature. The samples were analysed for the physico-chemical and biological parameters at different times, viz., zero, seven, fourteen, twenty one and twenty eight days. The values obtained are represented as a percent reduction in the values of physico-chemical and biological parameters.





CHAPTER - IV

RESULTS AND DISCUSSIONS

To determine the extent of pollution in river Dhamola, the various samples of water collected were analysed for three basic criteria to evaluate the quality of water. They are physical, chemical and biological. The physical and chemical parameters are concerned with a variety of procedures and in general depend on the source and many other factors. However, a complete chemical analysis of water provides a reasonably accurate picture of the quality of water. The biological analysis indicates about the pollution in terms of the presence of intestinal bacteria and other microscopic organisms in water.

#### 4.1 RIVER WATER OBSERVATIONS

The water samples in river Dhamola were collected for the period July 1988 to June 1990. The details and frequency of the collection have been described in Chapter - 3. The physico-chemical characteristics of the analysed samples are described below :

##### 4.1.1 Turbidity

The turbidity of the samples collected at various sampling stations of river Dhamola was found in the range of 18-130 NTU. It was however, observed that turbidity for all the three sampling stations was higher in the monsoon season [August, September, 1988] and then

decreased in the period of October 1988 to April 1989. It was interesting to note that the value of turbidity at sampling station 1 was always lower than stations 2 and 3 through out the year. The higher values of turbidity at all the three sampling stations in the monsoon season have been ascribed to the rainfall which increase the amount of suspended solids and thus causes an increase in turbidity. Similar observation has also been reported in the literature by Canale et al. [129]. A detailed seasonal variation of turbidity is presented in Fig. 4.1(a).

#### 4.1.2 Total Solids

The total solids (mg/l) determined at sampling station 1 was always found much lesser than at sampling stations 2 and 3 as prescribed in Fig. 4.1(b). The value of total solids during the rainy season (July - September, 1988) were always found higher at all the three sampling stations. This may be due to the presence of sandy particles coming from mountains as suggested by Valdiya and Bartarya [130]. A similar phenomenon has also been observed for Ganga water [131] which indicated that the maximum number of total solids was found during the rainy season, whereas, minimum was in winter and summer. Vander-Weijden et al. [62] have also observed the higher value of total solids during rainy season in Rhine river and explained it due to the deposition of suspended

matter in river. The total solids were also found to be practically similar at stations 2 and 3 during the period of April to June, 1989 and thus indicated that during this period the heavy particles present in river Dhamola do not settle for a stretch of about 5 kms. (distance between sampling station 2 and 3). Hence the increase in total solids due to the mixing of waste water/effluents at station - 2 remains same even after 5 kms of mixing at station 3. The higher value of total solids observed at sampling stations 2 and 3 was due to the mixing of waste water further finds support from the observation of Bendic and Stojanta [10] according to which municipal sewage and effluents have been reported to possess higher value of suspended particles.

#### 4.1.3 pH

The value of pH represents the  $H^+$  concentration for samples. The change of pH with seasonal variations for all the three sampling stations are presented in Fig. 4.1(c). It is observed that the pH value of river water at sampling station 1 varied in the pH range of 7.15 - 7.95. It was found lower than sampling stations 2 and 3 in all seasons except in the month of September, 1988, where the value exceeded to 7.95. One of the probable reasons for this increase in pH may be the heavy rains which were observed in the month of September 1988 (rainfall in September 1988, 139 mm). At sampling stations 2 and 3 the pH value of the samples was slightly



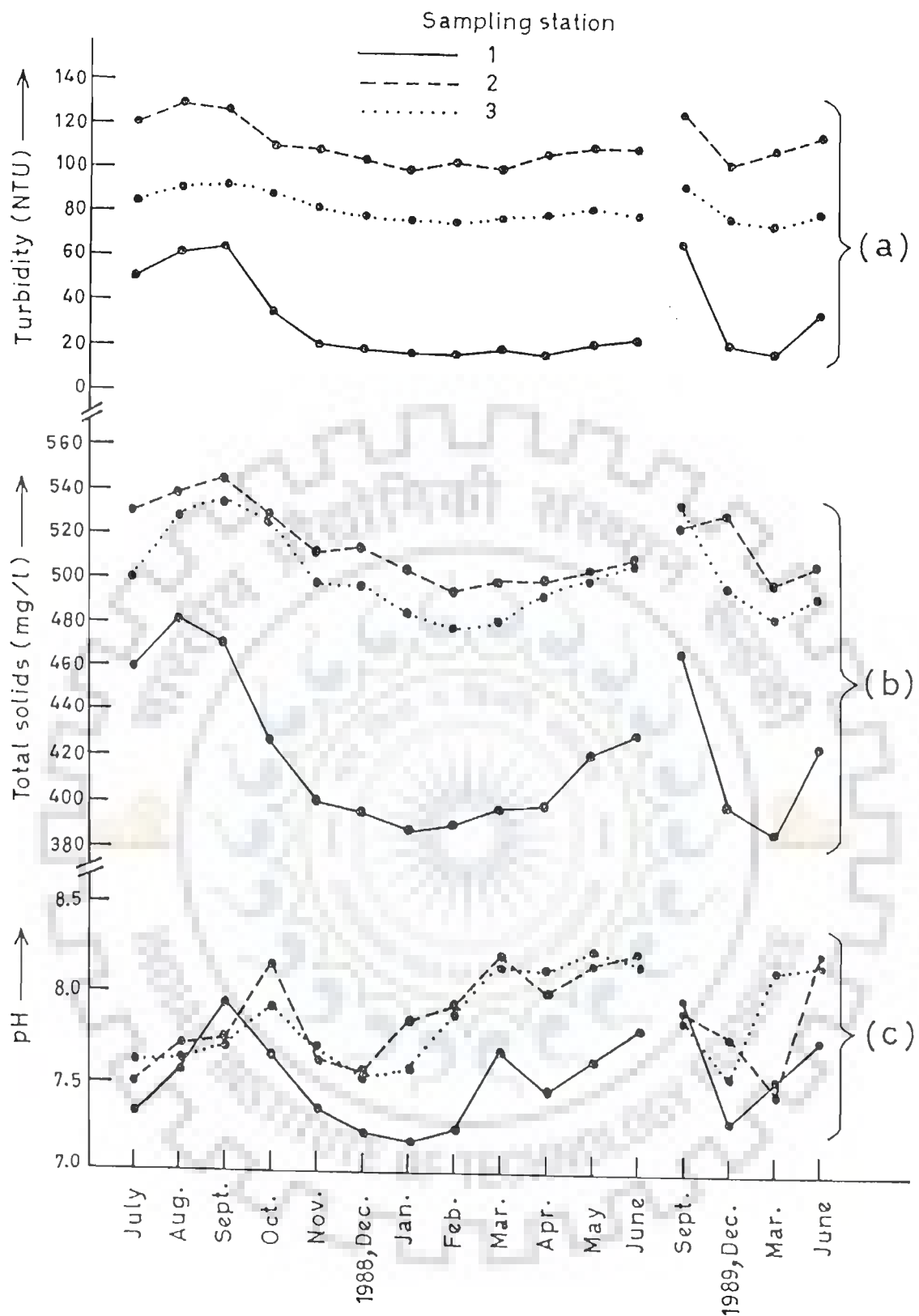


Fig. 4.1 Seasonal variations observed for turbidity, total solids and pH at different sampling stations from July 1988 to June 1990

higher in October 1988 and in the period of March - June 1989. During the winter season (November 1988 to February, 1989), the pH values were relatively lower at all the three sampling stations. The increase in the pH value of the river water during rainy season has also been suggested by Hutchinson [132] and Welch [133]. Such an increase has been suggested due to some oxidation - reduction reactions in the presence of degradable and non-degradable matter, during rainy season. The decrease in pH values during winter has been suggested due to the lowering of suspended matter and also due to the decrease in temperature [134]. The carbon dioxide equilibrium also causes a decrease in pH. It affects the water quality due to the changes in biological activity [135,136]. The variation in pH in 1989-90 also indicated practically similar trend.

#### 4.1.4 Total Hardness

The total hardness value of samples collected at sampling station 1 was found in the range of 129-151 mg/litre during the entire period of investigation and thus indicated that water at sampling station 1 is not hard in nature. The variations in the hardness at sampling station 1 with seasonal variation were not significant and are depicted in figure 4.2(a). However, the value of hardness observed at sampling stations 2 and 3 was much higher and practically identical at stations 2

and 3. Thus, after the mixing of waste water from sewage and industrial effluents, the hardness increased upto 249.0 mg/l. It was thus concluded that due to the mixing of effluents and sewage in the river Dhamola, there is a significant increase in hardness, which lasted atleast for a stretch of 5 kms (distance between points 2 and 3). The increase in total hardness upto about 350 ppm in Ganga and Bhagirathi river waters due to the mixing of industrial effluents has also been reported in the literature [137,138].

#### 4.1.5 Chloride

The value of chloride ion ( $\text{Cl}^-$ ) content in the water samples collected at sampling station 1, was always observed lower than sampling stations 2 and 3, throughout the period of this investigation. The higher value of  $\text{Cl}^-$  in the waste water stimulates the  $\text{Cl}^-$  concentration in running water [59]. The maximum value of chloride ( $\text{Cl}^-$ ) ions was found in the month of December 1988 and 1989 (24 mg/l and 21 mg/l respectively). The minimum value of chloride at same sampling station was found during September - October, 1988. Takakoe [11] on the other hand found maximum value of chloride ions in the month of September - October of 1977 in Nagase river in Japan. The value of chloride was found much higher at sampling stations 2 and 3 in comparison to sampling station 1 during the course of the investigation. Nevertheless, the value of chloride at all the three sampling stations

reached to a maximum during winter season. The detailed seasonal variations of chloride ion at all the three sampling stations are presented in Fig. 4.2(b).

#### 4.1.6 Free Carbon-dioxide

Fig. 4.2(c) depicts the effect of seasonal variation on the free  $\text{CO}_2$  of water samples at all the three sampling stations. The carbon dioxide content was found to change significantly with seasonal variations. At sampling station 1, the  $\text{CO}_2$  value showed a variation in the range of 0.2 - 1.1 mg/l during the entire investigation period. Samples from sampling stations 2 and 3 indicated an increase in  $\text{CO}_2$  value in all seasons except in the months of March, June and September, 1989. The higher value of  $\text{CO}_2$  at sampling stations 2 and 3 were probably due to the mixing of waste water from domestic sewage as well as industrial effluents at about 1 km upstream than sampling station No.2. After 5 kms downstream from the sampling station 2, the  $\text{CO}_2$  value slightly decreased but was still higher than upstream (station 1). The low value of  $\text{CO}_2$  observed at sampling station 1 is probably due to the great utilization of free  $\text{CO}_2$  during the photosynthesis process by the aquatic phytoplankton in fresh water, which liberates dissolved as well as free oxygen. Whereas, at sampling stations 2 and 3, the higher values of  $\text{CO}_2$  are due to the degradation of organic matter by the waste water discharge. Thus, it is anticipated that  $\text{CO}_2$  will exhibit

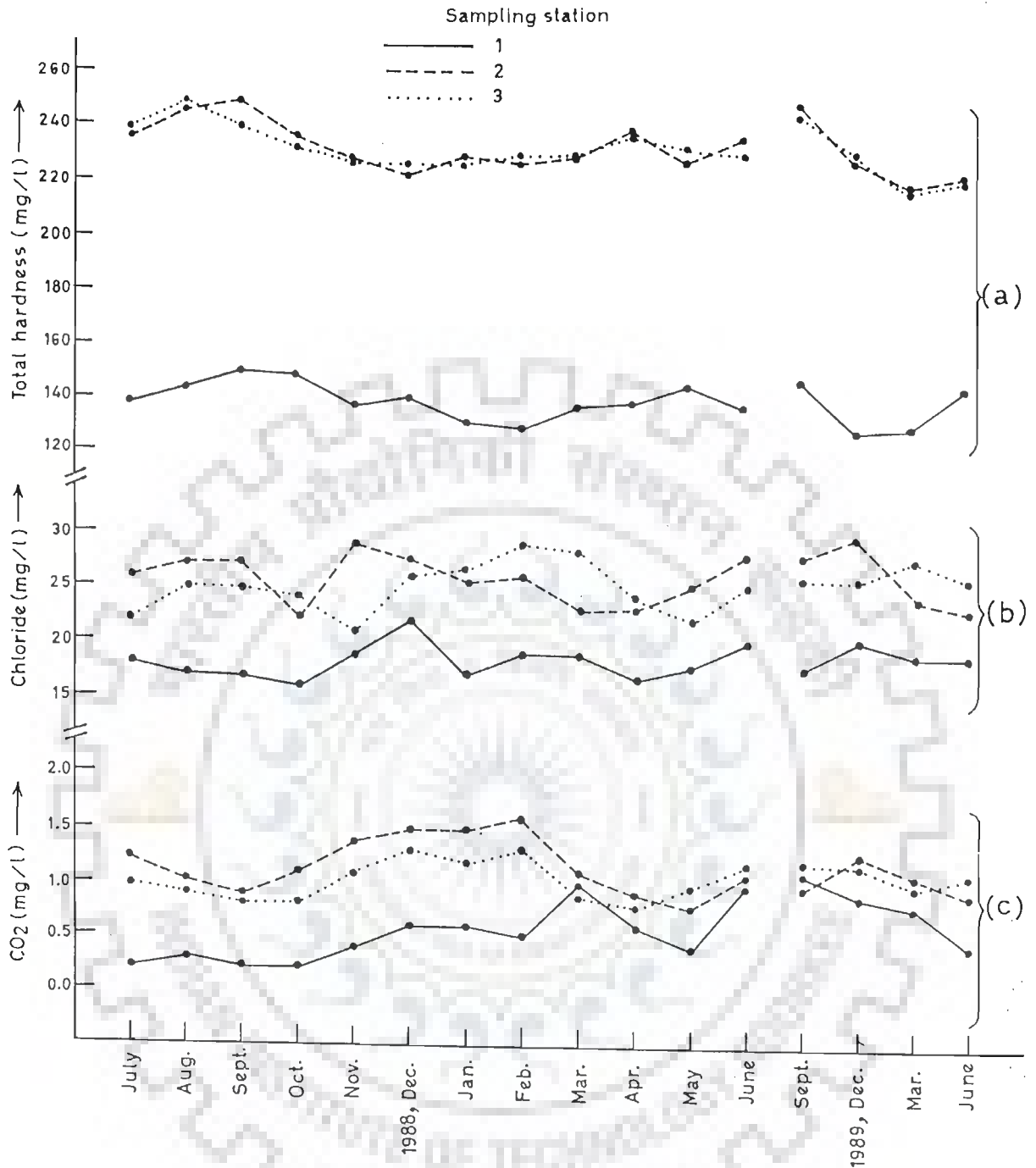


Fig. 4.2 Observed seasonal variations for total hardness, chloride ( $\text{Cl}^-$ ) and  $\text{CO}_2$  from July 1988 to June 1990

an inverse relationship with dissolved oxygen. Similar observation has also been reported in the literature by Shashi Kant and Raina [139] during limnological studies of two ponds in Jammu. Sinha et al. [140] has also observed a similar behaviour in Ganga water.

#### 4.1.7 Dissolved Oxygen (DO)

The variations in dissolved oxygen content with changes in season are presented in Fig. 4.3(a) which clearly indicates that the amount of dissolved oxygen at sampling station 1 is much higher than at sampling stations 2 and 3 during the entire period of investigation. The DO value observed at station 2, did not indicate a significant variation except in the months of June, September and December 1989 where the values observed were slightly higher. The high value of DO at station 1 can be assigned to the low organic degradable matter and high population of phytoplankton in the river Dhamola which provides a large amount of oxygen through photosynthesis [141]. The direct relationship of DO with phytoplankton, algae etc. has also been reported in the literature [142-144]. Sampling stations 2 and 3 indicated a lower value of DO probably due to the mixing of sewage and effluents and thus has been found very close to 5.0 mg/l. It thus indicated that the DO value at stations 2 and 3 is very close to the value required by aquatic life. The depletion in DO level after the mixing of waste water occurred probably due to the degradation of organic matter by

foreign particles of sewage as well as industrial waste. Such a depletion of DO by sewage water as well as industrial waste has been observed for several miles downstream by Alabaster [145].

#### 4.1.8 Biochemical Oxygen Demand (BOD)

Biochemical oxygen demand in fresh water of the river Dhamola was found to be influenced significantly by the mixing of waste water of domestic sewage and industrial effluents. At sampling station 1 (upstream), the BOD value was in the range of 8.0 - 18.0 mg/l throughout the period of investigations. The minimum BOD at point 1 was found during August and September 1989. The low value of BOD in upstream region of Ganga river during the period of August and September in 1980's has also been observed earlier by Sinha et al. [131] and has been accounted for on the basis of heavy rainfall. In August and September 1989, the northern part of India also observed a heavy rainfall (139 mm) and hence probably caused a decrease in BOD value. After the mixing of waste water, the BOD value increased to 86 mg/l at sampling station 2. The increase in the BOD value after the mixing of waste discharge usually occurs due to the deposition of organic degradable matter in river water and indicate the organic pollution load in the water [146]. Similar observation has also been reported by Seung et al. [147]. A 600% increase in BOD value due to mixing of industrial waste in Beronka river has also been observed earlier by

Vymazol [148]. Respass [149] has also shown the elevation in BOD value due to the mixing of sewage as well as effluents in river water. The BOD value at sampling stations 2 and 3 was almost identical during the period of July to October 1989 and then the value indicated a slight decrease at sampling station 3. The reduction in BOD value at sampling station 3, in comparison to sampling station 2, may be due to the self purification nature of sewage/effluent water [150]. The relation of BOD at all three stations with seasonal changes are presented in Fig. 4.3(b).

#### 4.1.9 Chemical Oxygen Demand (COD)

Like biochemical oxygen demand the chemical oxygen demand value of fresh water (sampling station 1, Fig. 3.2) was also found to increase after the mixing of sewage and industrial waste water. The value of COD at sampling station 1 was found in the range of 27-38 mg/l as presented in Fig. 4.3(c). In the months of August and September 1988, the values were lowest and in the months of May and June, highest. The low value of COD during the months of August and September, is usually observed due to the rainfall whereas, maximum value is observed in the summer season. Similar observations have also been reported in literature by large number of workers [131,134]. The COD value at this point was very low in comparison to the sampling stations 2 and 3. At sampling



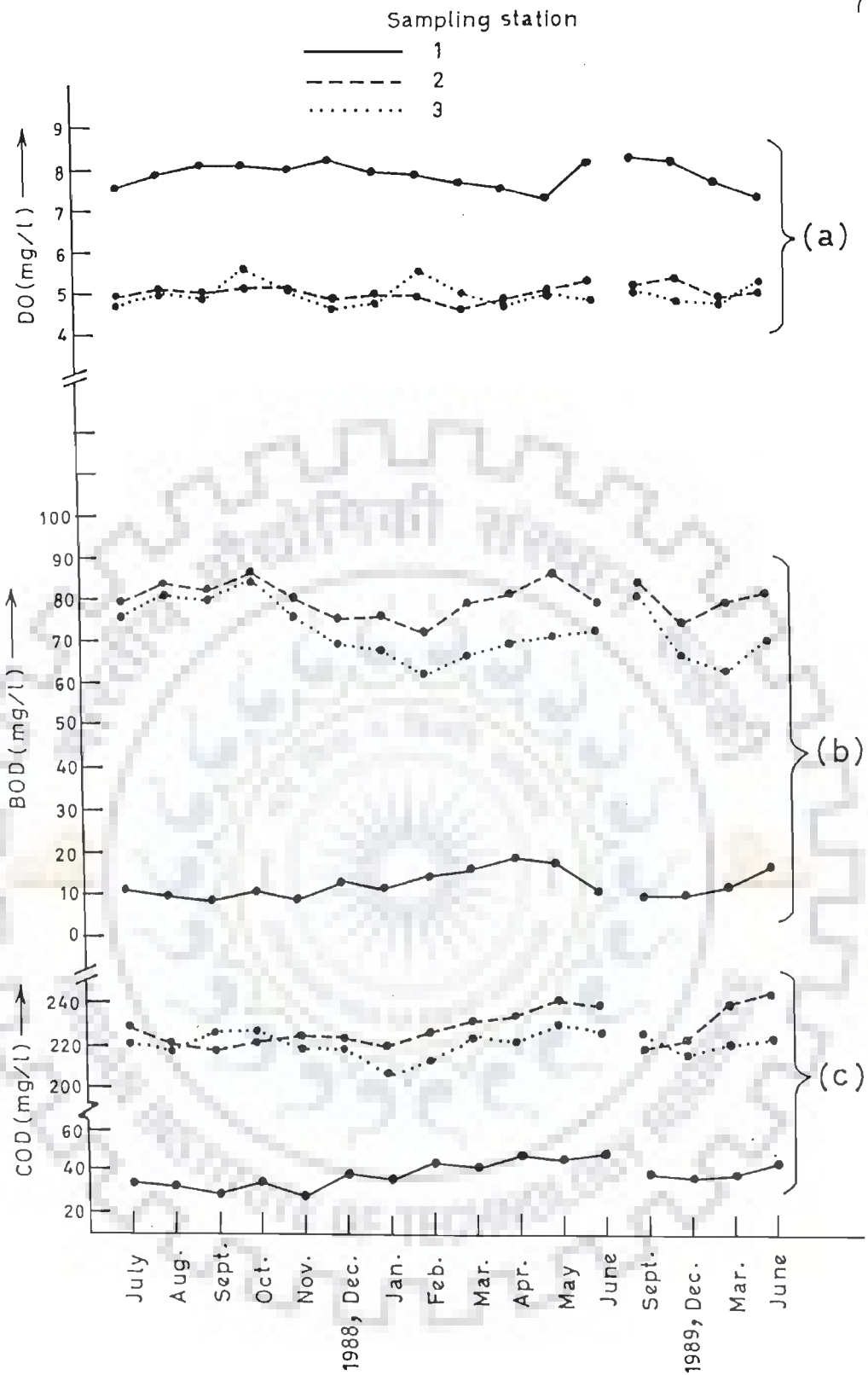


Fig. 4.3 Seasonal variations observed for DO, BOD and COD at different sampling stations

station 2, the value of COD was found in the range of 218-244 mg/l and was practically identical to sampling station 3 for the months of July to November 1988. Lalman and Dixit [138] and others [6] have also suggested the elevation in COD value of fresh water at downstream point due to the mixing of industrial waste disposal.

The use of river water with high value of COD may cause health hazards to both humans and domestic animals as suggested by Bermejo et al. [151]. Therefore, it was concluded that the water of river Dhamola should not be used for bathing purpose for mammals. After November, the COD value indicated a decreasing trend at sampling station 3 in comparison to station 2. The values of COD observed at all the three sampling stations are presented in Fig. 4.3(c).

## **4.2 BIOLOGICAL PARAMETERS**

### **4.2.1 Bacteriological Population**

The bacteriological characteristics determined for the various samples obtained for the river Dhamola are presented in Figs. 4.4 and 4.5.

#### **4.2.1.1 Most Probable Number of Coliform Bacteria**

The most probable numbers (MPN) of coliform bacteria at sampling station 1 (upstream) were found in the range of 23-460 per 100 ml in most of the months of

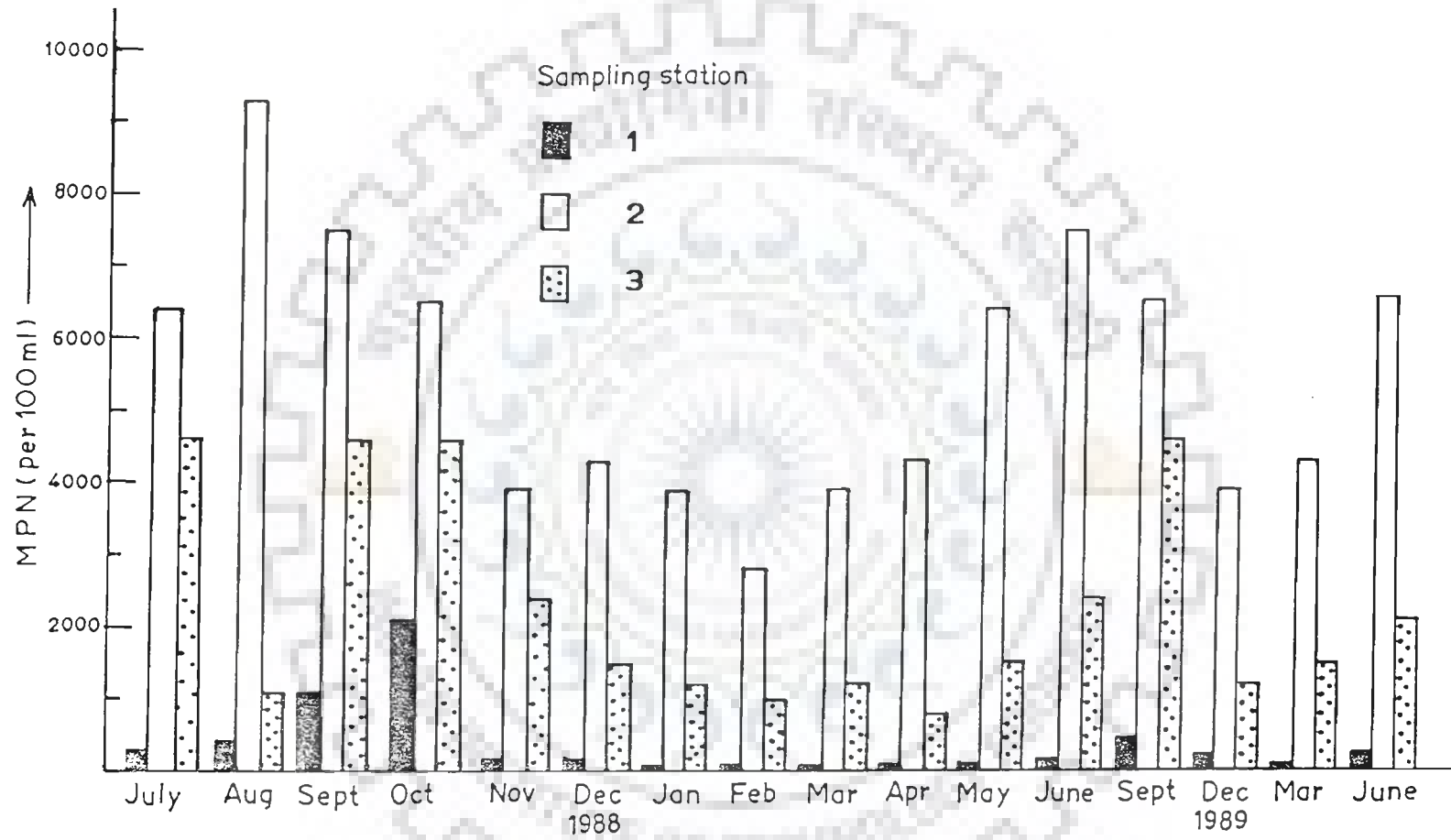


Fig. 4.4 Seasonal variations observed for MPN at sampling stations 1,2 and 3

investigations. An unusually high value (2100 per 100 ml) was however, observed in the month of October, 1988 probably due to the heavy rains observed in October 1988 in the northern part of India. The higher value of MPN during September - October, 1988 can be assigned to the heavy rainfall. This period has been claimed as the best survival period for coliform bacteria [129]. The MPN values at sampling stations 2 and 3 were much higher than at sampling station 1 during the entire course of the investigations and are presented in Fig. 4.4. It is thus concluded that mixing of municipal sewage and industrial effluents provided a medium suitable for the growth of bacteria, which caused a tremendous increase in MPN count. Much higher value of MPN at confluence point has also been observed earlier and is explained due to the mixing of sewage water as well as industrial effluents. The suspended organisms in waste water provide the nutrient level as well as organic matter for survival of coliform bacteria [6,152,153]. Dmitriew and Malyutin [12] have also suggested that the elevation in coliform bacteria occurs due to the mixing of waste water in river. This value, decreased at sampling station 3 most probably due to self-purification as reported in the literature [150].

#### **4.2.1.2 Standard Plate Count**

Like most probable number of coliform bacteria (MPN), the value of SPC also changed significantly with seasonal changes. All three sampling stations indicated a

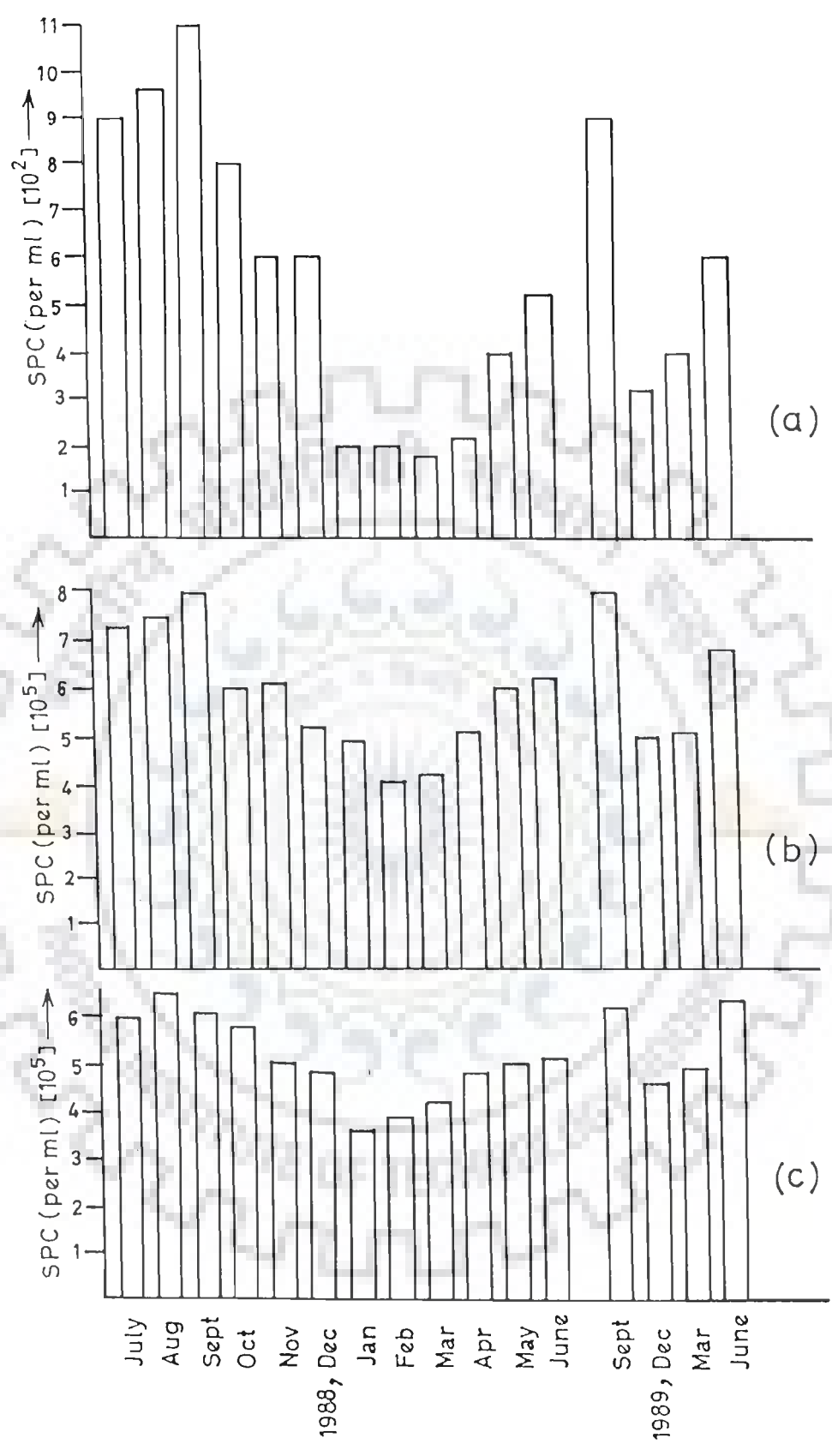


Fig. 4.5 Seasonal variations observed for SPC at sampling stations (a)<sub>1</sub> (b)<sub>2</sub> and (c)<sub>3</sub>

similar pattern of increase, decrease and increase in standard plate count of coliform bacteria due to the seasonal variations. The maximum number of SPC was found in the month of September 1988 at all the three sampling stations and minimum was found in the months of January - March 1989. The reduction in SPC during the months of January to March has also been reported by Canale et al.[129] as least survival period for total coliform bacteria. At sampling station 1, the SPC of coliform bacteria were very low ( $1.8 - 11.0 \times 10^2/\text{ml}$ ) as compared to the sampling stations 2 and 3 during the entire course of investigation. It is clear from the presented data that the value of SPC increased to about  $10^3$  times at sampling station 2, in comparison to station 1 due to the mixing of waste water. Nevertheless, the decrease in SPC value at station 3 was not significant. In the case of MPN, the value decreased to about 40 - 50% at station 3 in comparison to station 2, whereas in SPC, the value decreased to only about 10 - 20%. This difference in behaviour may be accounted for on the basis of selfpurification of sewage water as suggested by Pillai et al.[150].

#### 4.2.2 Planktonic Population

The various plankton species identified in the water samples collected from the different sampling stations are presented in table 4.1. It can be clearly seen that several pathogenic protozoans were identified only at

Table 4.1 List of different species of algae, diatoms and protozoans identified in the water samples collected from different sampling stations

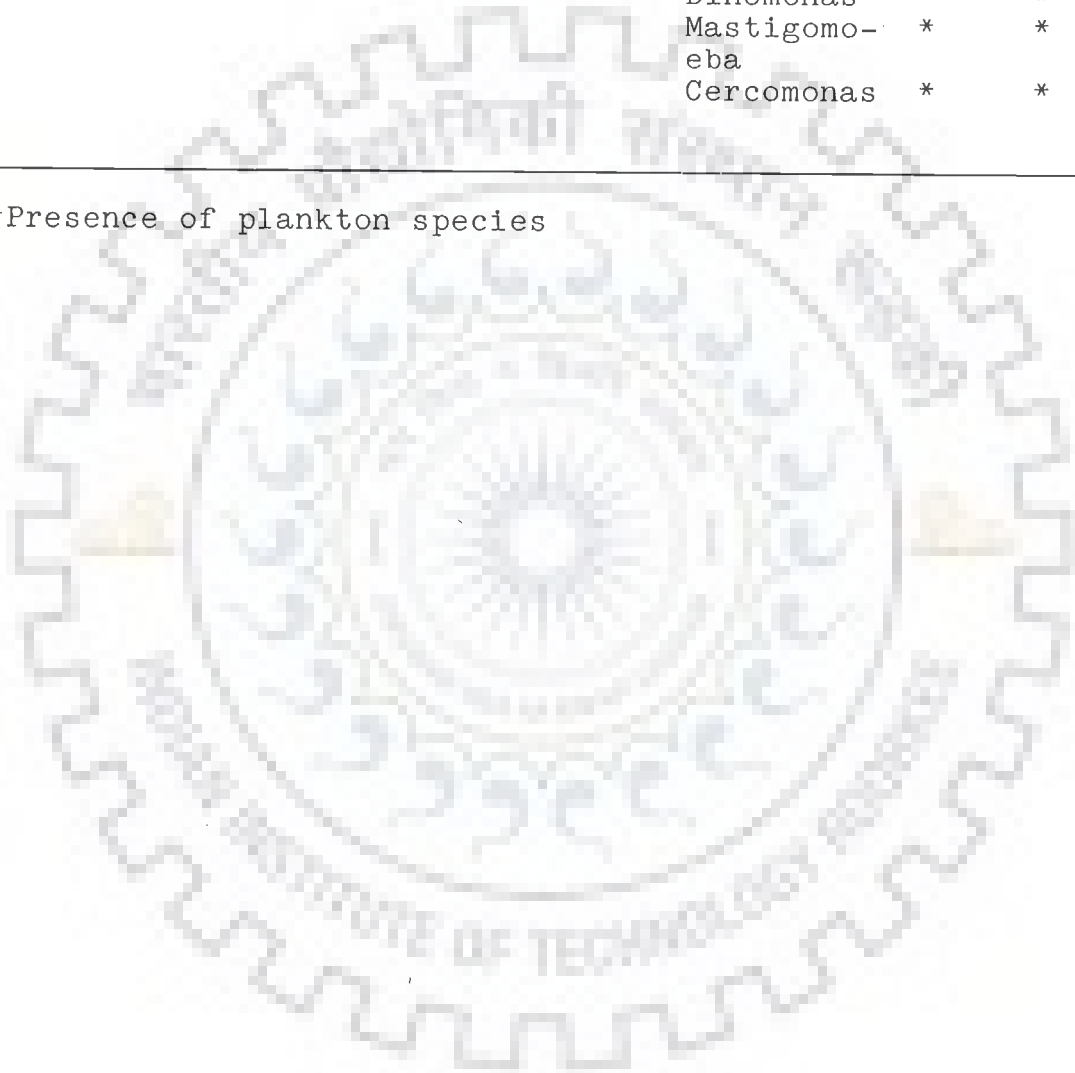
Parameters	Phyllum	Nature	Plankton species/ genus	Sampling Stations				
				1	2	3	4	5
Algae	Chlorophyta	Filamentus	Pithophora	*				
			Schizameris	*				
			Zygnema	*	*	*		
			Spirogyra		*	*	*	
			Hylothecca	*				
			Chladophora	*				
		Coccoid	Dictyosphaerium	*				
			Westella	*				
			Chlorella	*				
			Crucigenia		*	*	*	
	Chlamydomonas			*	*	*		
	Cyanophyta	Filamentus	Occillatoria		*	*	*	
			Pinnularia	*		*		
			Lyngbya		*	*		
			Anabaena		*	*	*	
Gleotrichia			*		*			
Coccoid			Anacystis		*	*		
Diatoms		Chrysophyta	Pennate	Diatoma		*	*	*
	Fragilaria			*	*	*		
	Synedra				*	*	*	
	Achnanthes				*	*	*	
	Navicula			*		*		
	Surirella			*		*		
	Centric		Cyclotella	*		*		
			Coscinodiscus		*	*		
			Melosira		*	*	*	
Protozoans	Protozoa	Amoebas	Amoeba sp.	*	*	*		
			Diffflugia	*	*	*		
			Arcella	*	*	*		

Contd..... /-

Ciliates	Pleuronema	*	*	*
	Paramecium	*	*	*
	Coepoda	*	*	*
	Euplotes	*	*	*
	Aspidisca	*	*	*
	Vorticella	*	*	*
	Euglina	*	*	*
Non-pigmented	Peranema	*	*	
	Astaria	*	*	
	Dinomonas	*	*	
	Mastigomoeba	*	*	
	Cercomonas	*	*	

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\*Presence of plankton species





stations 2 to 4 during the entire course of investigation and thus indicate that the water of river Dhamola is polluted at stations 2 and 3. The most probable cause of such a pollution appears to be mixing of municipal sewage (station 4), which also indicated a positive test for these pathogenic protozoans. Similarly several species of diatoms were identified only in water samples obtained from station 1, whereas stations 2 to 5 did not indicate their presence. Thus mixing of sewage waste and industrial effluents kills the diatoms of fresh water.

#### 4.2.2.1 Algae

Fig. 4.6(a) represents an interesting alterations in algal count (primary producer) of the water of river Dhamola, after the mixing of waste water. The algal population varied with the change in season and varied in the range of 1050 - 1500 per ml for station 1. The maximum number of algal population was observed in the months of September, October, 1988 and June 1989, 1990. The increase in the algal count in these months is probably due to the dilution of water caused by heavy rains. Fogg [154] and Hutchinson et al. [155] have also suggested the presence of maximum number of algae in the rainy seasons as this period provides the required amount of nutrients. The value of algal count decreased to about five to six times at station 2 and 3 in the entire seasons of investigations. The observed reduction in algal count after the mixing of waste water is due to the influence of

temperature and excess of chloride ( $\text{Cl}^-$ ) in the waste water [155,156]. Hutchinson and Stokes [155] have also suggested the low value of algae after mixing of municipal sewage due to the increase of organic and inorganic content. The algal population at stations 2 and 3 were essentially same and thus it was concluded that the algal population remained the same for the stretch of about 5 kms after mixing of the domestic sewage and industrial effluents. This behaviour is in conformity with the findings of Buzzel et al. [157] and Joseph et al [158].

#### 4.2.2.2 Diatoms

The observed variations of total number of diatoms per millilitre during the period of two years is presented in Fig. 4.6(b). It indicated a very significant change in the population of diatoms in the water of river Dhamola after mixing of domestic sewage and industrial effluents. The value of diatoms at sampling station 1 was practically constant (in the range of 50-80/ml) during the entire course of investigations. The value increased to about 18-25 times at sampling station 2. The increase in the value of diatoms at sampling station 2 is due to the mixing of organic matter from the industrial effluents, which provides nutrients for the growth of diatoms as suggested by Patrick [159]. The diatoms value then decreased at sampling station 3. The highest value of diatoms at stations 2 and 3 was observed in the month of

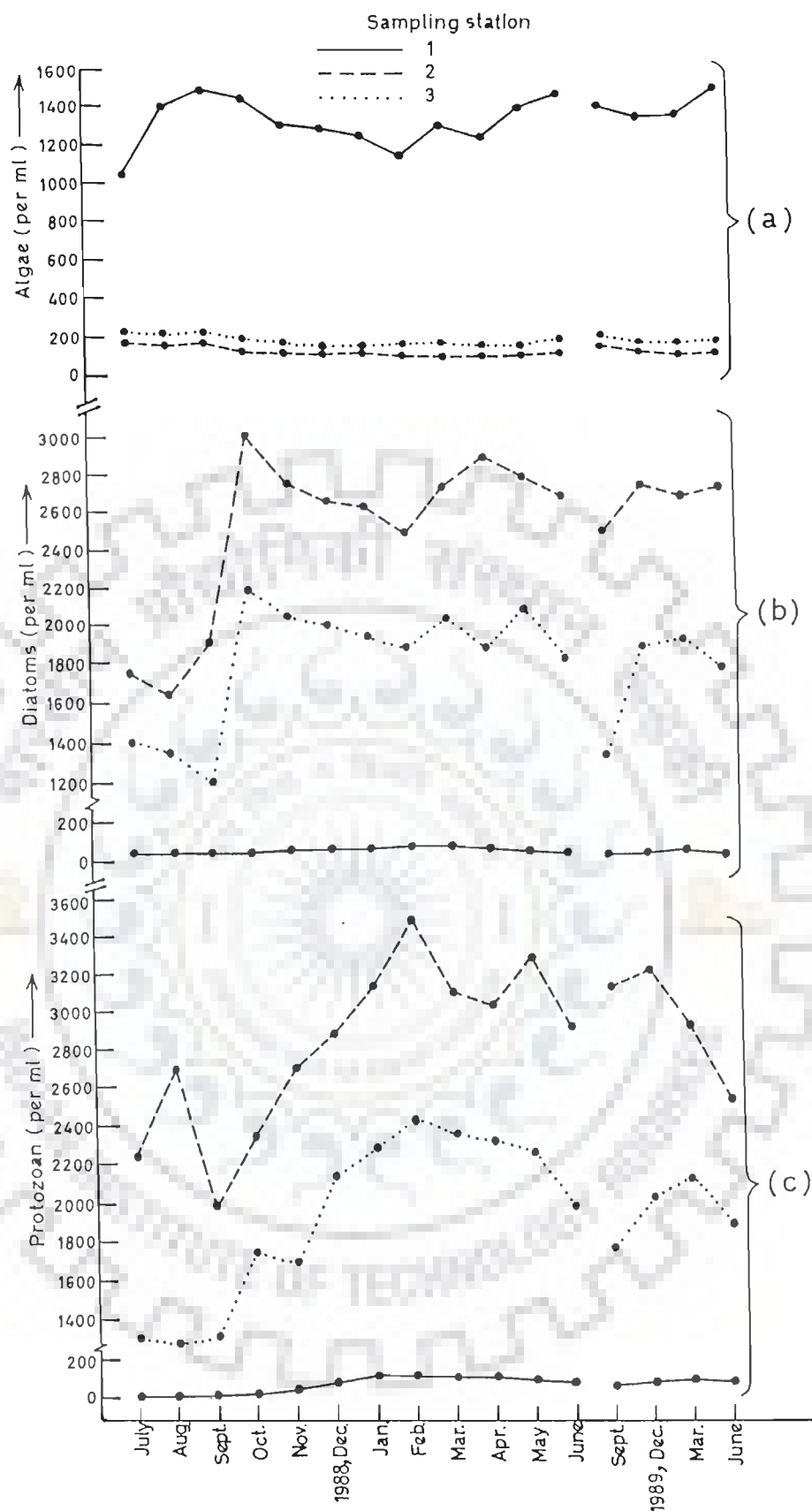


Fig. 4.6 Seasonal variations observed for (a) algae, (b) diatoms and (c) protozoans from July 1988 to June 1990

October 1988. This behaviour may be accounted for on the basis of increase in BOD value in the month of October 1988 at sampling stations 2 and 3, which caused the increase in diatoms value. Similar observations have also been reported in the literature for Periyar river [153]

#### 4.2.2.3 Protozoans

Fig. 4.6(c) presents the change in the population of protozoans with seasonal variations. The value of protozoans at station 1, was almost constant and did not change with seasonal fluctuations. However, this value increased significantly after mixing of municipal sewage and effluents. Thus at station 2, the value of protozoans did not show any regular pattern but had an increasing trend from September 1988 to February 1989. The maximum values of protozoans were observed in the month of February 1989 at both the stations 2 and 3 (3500 and 2440 per ml, respectively). The elevation in protozoans population after mixing the municipal sewage has also been observed earlier [6,160].

### 4.3 WASTE WATER CHARACTERISTICS

The analysis of waste water is important because it provides information about the nature and extent of pollution load present in it. Hence, it becomes easier to find ways and means to reduce the pollution in the river in which it mixes. In the present investigation the physico-chemical and biological analysis of waste water

from domestic sewage (sampling station 4, Fig. 3.2) and industrial effluents (station 5, Fig. 3.2) mixing in the river Dhamola were carried out during the period of July 1988 to June 1990. The method for the collection of samples and their frequency were essentially the same as that for the river water samples and have been described in Chapter 3. The results obtained for the various parameters are summarised in Table 4.2 for the domestic sewage and in Table 4.3 for industrial effluents.

It is clear from the Table 4.2 that all the physico-chemical parameters for the samples of domestic sewage collected during the entire period of investigation were much higher than fresh water parameters except the dissolved oxygen which was found very low (1.6-2.6 mg/l). The value of DO was thus very less than the minimum permissible limits [161] of Indian Standard as well as of World Health Organisation (5.0 mg/l). The low value of DO throughout the period of two years in the domestic sewage indicated the presence of high pollution load, which was further confirmed by the high value of BOD, COD and other physico-chemical parameters. The low value of DO in domestic sewage has also been observed earlier [157] and assigned to the high value of BOD, COD and absence of phytoplankton (primary produce).

All biological parameters were also found to be very high in the samples of domestic sewage (Station 4) except the algal population which indicated a decreasing trend.

Table 4.2 Observed physico-chemical and biological characteristics of domestic sewage at sampling station - 4.

Year Month Parameter	1988						1989						1990			
	July	Aug.	Sept	Oct	Nov	Dec	Jan	Feb	March	Apr	May	June	Sept	Dec	March	June
Turbidity (NTU)	295 ±5	297 ±5	290 ±4	270 ±6	255 ±4	242 ±4	207 ±5	201 ±7	266 ±4	276 ±3	269 ±4	277 ±5	276 ±5	235 ±4	269 ±5	270 ±3
Total sol- ids(mg/l)	875 ±7	860 ±7	880 ±7	810 ±7	774 ±4	766 ±5	750 ±5	714 ±4	695 ±8	725 ±7	717 ±6	763 ±7	720 ±6	780 ±8	711 ±8	715 ±7
pH	7.99 ±0.1	8.10 ±0.1	8.00 ±0.1	8.09 ±0.1	8.17 ±0.1	8.01 ±0.1	8.03 ±0.1	8.07 ±0.0	8.04 ±0.0	8.13 ±0.1	8.12 ±0.1	8.04 ±0.2	8.10 ±0.1	8.00 ±0.1	8.10 ±0.1	8.10 ±0.0
Total hard- ness(mg/l)	271 ±3	283 ±3	278 ±2	269 ±2	362 ±2	371 ±2	366 ±3	369 ±4	273 ±2	271 ±3	269 ±2	275 ±1	280 ±1	380 ±2	282 ±1	274 ±2
Chloride (mg/l)	30 ±0.1	31 ±1.1	30 ±1.2	29 ±1.0	28 ±1.2	28 ±1.0	29 ±1.0	29 ±1.0	30 ±1.0	28 ±1.0	28 ±1.0	28 ±0.9	32 ±1.0	28 ±0.9	31 ±1.9	30 ±1.0
CO <sub>2</sub> (mg/l)	2.6 ±0.0	2.5 ±0.1	2.3 ±0.1	2.0 ±0.1	2.9 ±0.1	2.2 ±0.1	2.0 ±0.1	2.1 ±0.2	2.9 ±0.1	2.7 ±0.1	2.8 ±0.1	2.9 ±0.1	3.0 ±0.1	1.9 ±0.1	3.1 ±0.1	3.0 ±0.1
DO(mg/l)	2.6 ±0.2	2.4 ±0.2	2.6 ±0.2	1.9 ±0.3	2.4 ±0.3	2.3 ±0.4	1.7 ±0.2	1.6 ±0.2	1.9 ±0.2	1.7 ±0.3	1.7 ±0.3	1.8 ±0.3	2.5 ±0.3	2.5 ±0.3	2.0 ±0.2	2.2 ±0.2
BOD(mg/l)	175 ±2	210 ±2	217 ±2	231 ±	209 ±2	213 ±2	231 ±2	237 ±2	243 ±3	234 ±3	235 ±3	231 ±3	240 ±3	205 ±2	251 ±2	232 ±2

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COD (mg/l)	209 ±2	227 ±2	224 ±2	240 ±1	225 ±2	237 ±2	260 ±2	259 ±3	267 ±2	250 ±1	262 ±2	249 ±3	245 ±1	234 ±2	259 ±2	239 ±3
MPN x 10 <sup>4</sup> (per 100 ml)	25 ±0.3	23 ±0.4	19 ±0.4	20 ±0.3	18 ±0.2	15 ±0.2	16 ±0.2	19 ±0.3	20 ±0.2	23 ±0.6	18 ±1.0	14 ±0.8	17 ±0.2	15 ±0.3	18 ±1.0	12 ±1.0
SPC x 10 <sup>5</sup> (per ml)	28 ±1.0	28 ±1.3	28 ±1.0	27 ±1.0	28 ±1.0	26 ±1.4	27 ±1.3	28 ±2.0	26 ±1.0	23 ±1.0	24 ±1.0	22 ±1.0	23 ±1.0	25 ±1.0	24 ±1.0	20 ±1.0
Algae (per ml)	75 ±6	60 ±5	65 ±5	10 ±4	-	-	5 ±3	-	-	-	-	-	20 ±4	25 ±4	-	-
Diatoms (per ml)	2500 ±15	2400 ±16	2590 ±14	3500 ±21	3700 ±18	4300 ±21	4250 ±19	4100 ±20	4000 ±18	4400 ±18	4350 ±19	4450 ±20	2700 ±18	4200 ±17	3900 ±19	4300 ±17
Protozoans (per ml)	4000 ±23	4150 ±23	4050 ±24	4400 ±19	4750 ±18	5300 ±21	5100 ±21	5250 ±21	5700 ±22	5600 ±21	5750 ±21	5890 ±21	4500 ±23	5500 ±19	5800 ±20	5900 ±21

\* All values reported are an average of atleast two replicate determinations

**Table 4.3** Observed physico-chemical and biological characteristics of industrial effluent at sampling station 5

Year	1988						1989						1990			
Month	July	Aug	Sept	Oct	Nov	Dec	Jan	Feb	March	Apr	May	June	Sept	Dec	March	June
Turbidity (NTU)	336 ±4	358 ±4	349 ±5	330 ±5	332 ±5	317 ±5	316 ±5	309 ±5	319 ±4	330 ±3	337 ±4	341 ±3	356 ±4	348 ±4	325 ±5	339 ±4
Total Solids (mg/l)	1105 ±9	1140 ±9	1095 ±7	989 ±8	992 ±8	1002 ±10	978 ±9	999 ±6	1040 ±8	1080 ±7	1043 ±7	969 ±8	1125 ±10	992 ±8	1025 ±9	999 ±7
pH	8.39 ±0.3	8.47 ±0.2	8.43 ±0.1	8.24 ±0.3	8.06 ±0.3	7.93 ±0.3	7.97 ±0.2	7.95 ±0.2	8.13 ±0.1	8.01 ±0.2	8.19 ±0.1	8.23 ±0.1	8.56 ±0.1	7.91 ±0.2	8.22 ±0.2	8.05 ±0.2
Total hardness (mg/l)	310 ±4	313 ±5	302 ±5	320 ±4	339 ±6	326 ±5	329 ±5	319 ±5	307 ±4	319 ±4	325 ±4	328 ±6	296 ±5	346 ±6	315 ±6	339 ±6
Chloride (mg/l)	28 ±1	26 ±1	24 ±1	27 ±1	26 ±1	27 ±1	29 ±1	28 ±2	30 ±2	29 ±2	28 ±2	28 ±1	22 ±2	28 ±3	29 ±3	26 ±2
CO <sub>2</sub> (mg/l)	3.4 ±0.0	3.2 ±0.3	3.1 ±0.2	3.2 ±0.2	3.8 ±0.6	3.9 ±0.7	3.5 ±0.6	3.4 ±0.4	3.2 ±0.6	3.9 ±0.5	4.1 ±0.6	4.3 ±0.3	2.9 ±0.2	4.2 ±0.4	3.1 ±1.0	4.2 ±1.0
DO (mg/l)	1.3 ±0.0	1.4 ±0.1	1.6 ±0.2	1.5 ±0.1	1.0 ±0.3	0.8 ±0.3	0.8 ±0.2	2.0 ±0.3	1.7 ±0.3	1.6 ±0.2	1.4 ±0.3	1.2 ±0.3	1.8 ±0.3	0.9 ±0.2	1.9 ±0.4	1.4 ±0.2

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BOD (mg/l)	302 +3	290 +7	276 +4	299 +10	343 +9	329 +8	331 +8	319 +6	301 +9	293 +10	298 +9	317 +9	268 +10	302 +8	279 +9	296 +6
COD(mg/l)	338 +9	322 +8	309 +10	367 +9	447 +9	423 +9	468 +8	417 +9	399 +6	349 +5	360 +6	367 +8	299 +8	401 +7	385 +9	376 +8
MPN x 10 <sup>4</sup> (per 100 ml)	27 +5	26 +7	26 +6	23 +5	21 +6	25 +3	29 +5	26 +3	27 +6	25 +5	26 +4	26 +4	24 +6	22 +6	23 +5	22 +5
SPC x 10 <sup>5</sup> (per ml)	36 +1	38 +3	33 +2	37 +2	40 +2	44 +3	43 +3	43 +2	39 +2	34 +2	38 +4	42 +4	31 +4	41 +3	34 +2	37 +3
Algae (per ml)	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-
Diatoms (per ml)	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-
Protozoans (per ml)	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-

\*All values reported are an average of atleast two replicate determination.

The number of algal population was practically zero, except in the months of rainy season. It appears that insignificant value of algae is due to the high values of BOD and COD in the domestic sewage. It seems reasonable to conclude that the zero population of algae caused the low value of DO observed in the samples. The presence of MPN, SPC, diatoms and protozoans also indicate the high rate of pollution in the domestic sewage.

In the case of samples obtained from sampling station 5 (industrial effluents), the values of all the physico-chemical parameters, except DO were found higher than the permissible limit to discharge of waste water into inland surface water by Indian Standards [162]. The value of DO was found in the range of 0.5 - 2.0 mg/l during the entire period of investigation. The higher values of physico-chemical parameters indicated the high pollution load in the industrial effluents. The high values of BOD and COD pointed the higher organic pollution load due to the presence of organic matter whereas, other parameters indicated that the pollution is mostly due to the presence of inorganic as well as metallic ions.

The low value of DO and high values of BOD and COD observed in the industrial effluents during the entire period of investigation did not permit the growth of plankton populations. Hence, the total planktonic population was found to be practically zero. The absence of planktonic population in industrial effluents is well

known and has been assigned to the presence of organic and inorganic pollutants present in it. Area receiving the waste water discharge (pollution impact zones) has been reported to be completely devoid of all life except bacterial population at down stream of effluent [7]. Only MPN and SPC were observed in the samples collected at this sampling station. Such a behaviour is usual phenomenon as minimum permissible value of DO for the existence of aquatic life is 5.0 mg/l.

In general, it is observed that the values of all the physico-chemical parameters except DO were greater in industrial effluents than in the domestic sewage samples whereas, the value of DO in the industrial effluent was much lower than in the domestic sewage during the entire course of investigations. The plankton population was totally absent in industrial effluent. The MPN and SPC values were also higher in the industrial effluent than in the sewage water. The lower value of BOD in domestic sewage in comparison to the industrial effluents can be assigned to the excess of chloride ions in the domestic sewage and is in confirmity with the observation of Santhgate [163] and Snow [164] who have assessed that the chlorination of the domestic sewage and industrial effluents caused a sharp decrease in the BOD value.

#### 4.4 OXYGEN SAG

The dependence of DO was also monitored against the down stream distance from the mixing point of waste water in the river Dhamola to get an idea about the nature of oxygen sag curve. The curve obtained for the dependence of DO is presented in Fig. 4.7. It was thus observed that if the net influx of oxygen from the atmosphere to the river Dhamola at sampling station 1, is considered as 100% saturation, there is a considerable decrease in DO after the mixing of waste water. It is reported that if BOD loading from the waste water is sufficiently large the oxygen concentration in the river/stream may drop to virtually zero and may remain same for a considerable distance [150]. In the present studies, the DO of water decreased to about 50% after the mixing of the municipal and the industrial effluents. After about 5 kms of the mixing point, the DO value further indicated an increase to about 75%. Thus, it appears that the effects of reaeration and declining respiration rate allow the oxygen concentration to build up again during the stretch of 5 kms. It is well known [150] that the zone of anoxic water may extend to many kms before the oxygen concentration reaches to 100% again. Thus it could be concluded that it will be further several kms down stream from station 3, where the oxygen concentration build up to 100% again.



Fig. 4.7 Observed Oxygen Sag for the river Dhamola

It was considered necessary to compare the pollution load of the municipal sewage and industrial effluents mixing in the river Dhamola with the physico-chemical characteristics suggested by Indian Standard Institute [162] for waste water before allowing it to mix in the river. A comparison of four important characteristics, viz., pH, total solids, BOD and COD of municipal sewage, industrial effluent and suggested data by ISI is presented in Table 4.4. It was found that the values of pH of both the waste waters were in the range suggested by ISI (Indian Standard Institute) during the entire course of investigation. However, the values of BOD and COD were found much higher than the values recommended by ISI. Therefore, it is highly desirable that the waste water should be properly treated before allowing it to pass into the river Dhamola. If the waste water with such a high values for BOD and COD is allowed to mix (as it is happening in the present case) in the river, then it is expected that it would kill the aquatic life except bacterial population as suggested by Dickman et al.[7].

The physico-chemical and bacteriological parameters of the water samples collected at sampling stations 1 to 3 have also been compared with the ISI standards suggested for the domestic and agricultural uses of the water and are presented in Table 4.5.

**Table 4.4** A comparison of some important characteristics of domestic waste and industrial effluent mixing in the river Dhamola with the ISI suggested values[165)

Parameters	ISI (max tolerable value)	Domestic waste	Industrial effluent
pH	5.5 - 9.0	7.99 - 8.17	7.91 - 8.56
Total solids (mg/l)	2500	695 - 880	968 - 1140
BOD (mg/l)	20 - 30	175 - 243	276 - 343
COD(mg/l)	250	209 - 267	299 - 468

**Table 4.5** A comparison of observed value of physico-chemical and biological parameters for Dhamola river at sampling stations 1,2 & 3 during 1988-90 with ISI value of water quality parameters

Parameter	Indian Standard			Sampling Station 1	Sampling station 2	Sampling station 3
	Bathing	Drinking	Agriculture			
Turbidity (NTU)	30 (max)	25 (max)	-	18-64	100-129	77-96
Total Solids (mg/l)	600 (max)	500 (max)	2100 (max)	389-489	496-546	469-537
pH	6-9	6-9	5.5-9.0	7.18-7.97	7.52-8.23	7.56-8.24
Chloride (mg/l)	600 (max)	-	600 (max)	16-22	22-30	21-29
CO <sub>2</sub> (mg/l)	0.8 (max)	-	-	0.2-1.1	0.8-1.6	0.8-1.3
DO (mg/l)	5.0 or 40% Saturation	5.0 (min)	3.0	7.2-8.2	4.6-5.3	4.6-5.6
BOD (mg/l)	30 (max)	3-6	100 (max)	8.2-20.5	72-86	62-85
COD (mg/l)	50 (max)	Not more than 10	250	28-49	217-244	205-229
MPN (per 100 ml)	Not more than 100	Not more than 10	-	23-460	28000-93000	7500-46000
SPC (per ml)	Not more than 200	-	-	18-110	41000-80000	36000-65000



All the parameters at upstream sites of river Dhamola (sampling station 1) were found well within the range suggested by ISI for bathing and agriculture purposes [165]. The higher value of turbidity (64 NTU) was observed only during rainy season due to criping of upper soil layer from mountains. The value of COD was also about 2-3 times higher than recommended by ISI and therefore, it is concluded that the water of river Dhamola is not suitable for drinking even at sampling station 1. However, the physico-chemical and biological parameters at sampling station 1 clearly indicates that it can be used safely for bathing and agricultural purposes.

After the mixing of waste water from domestic sewage and industrial effluents, all the parameters altered significantly and indicated that the water quality of Dhamola river reaches to disbalanced conditions. The higher values of BOD, COD, CO<sub>2</sub> chloride ions, hardness and coliform bacteria than the recommended standard values suggests that the water of river Dhamola at middle and down streams is polluted and is not suitable for domestic use. Its use in bathing may cause some dangerous cronic diseases in human beings as well as in the domestic animals. The value of BOD and MPN are much beyond the permissible limits for water intended for even bathing at sampling stations 2 and 3 and hence the water should be used only for agricultural purposes.

#### 4.5 REMOVAL OF POLLUTION LOAD

Removal of pollution load, created by municipal sewage and industrial effluents is one of the major tasks to be handled by the appropriate agencies to control the disastrous results of pollution in future. The report of the Central Pollution Board, New Delhi indicating that only 30% of the industries along the banks of the Ganga and its tributaries are following some kind of pollution control measures, emphasizes the need of strict measures to be taken to control water pollution. Though, nature has its own system of checks and balances of replenishing the sources of air, water, vegetation and animal life, still the exploitation of nature's gifts should be stopped. In recent years, it has been found that several plants are capable of reducing toxic elements, like arsenic, cadmium, lead, zinc, vanadium, iron, mercury, chromium and other chemicals. The use of water hyacinth lagoon and Azolla, Valessneria, Lemna minor etc. for removing toxic elements in the river and lake water has recently been reported in the literature [94,98-101,116, 117,166-168]. Though, only inorganic and organic pollution studies have been carried out in the present investigations, it was considered interesting to use some of the plant species to find out if organic and inorganic pollution load is also decreased by these species. Very few attempts have been made in this direction and little information is available in the literature[169].

In the present study, three different aquatic plant species, viz., E. crassipes, L. minor and C. demorsum have been employed for the removal of pollution load, created by domestic and industrial effluents. All the physico-chemical, bacteriological and biological parameters e.g., turbidity, total solids, pH, total hardness, chloride, DO biochemical oxygen demand (BOD), chemical oxygen demand (COD), carbon dioxide (CO<sub>2</sub>), most probable number (MPN) and standard plate count (SPC) of coliform bacteria were determined at an interval of one week and showed significant change. The water quality of waste water had improved after about 4 weeks of the treatment. The experiments divided into five categories have been described in the materials and methods section of the thesis.

#### 4.5.1. Stage I

The physico-chemical, bacteriological and biological parameters obtained for Eichhornia crassipes culture are given in Table 4.6. It was observed that in the case of domestic sewage, the value of DO increased to about two times after 4 weeks. The maximum value of DO 214%, 5.6 mg/l was found in the month of June 1990 and is close to the minimum permissible value suggested by WHO and Indian Standard for aquatic life [161,165]. BOD, COD and free CO<sub>2</sub> decreased in comparison to the control values. The values of BOD and COD reduced to about 55% and 57%

Table 4.6 Observed values of various parameters for the stage I treated domestic sewage and industrial effluent

Parameters	Month	Domestic sewage										Industrial effluent									
		0 day		7 days		14 days		21 days		28 days		0 day		7 days		14 days		21 days		28 days	
		C	T	C	T	C	T	C	T	C	T	C	T	C	T	C	T	C	T	C	T
Turbidity	Sept	100	100	22	16	18	13	12	6.5	9	5	100	100	19	15	15	11	10	6	10	4
	Dec	100	100	26	19	22	14	15	8	13	6	100	100	19	15	15	11	10	5	10	4
	March	100	100	25	18	19	15	14	8	12	6	100	100	23	15	19	13	12	7	11	5
	June	100	100	20	15	17	12	14	7	11	5	100	100	20	15	15	12	10	6	11	5
Total solids	Sept	100	100	56	55	53	42	25	10	23	9	100	100	72	72	45	36	19	7	16	6
	Dec	100	100	66	66	56	49	25	10	22	9	100	100	83	83	50	40	21	8	17	7
	March	100	100	56	54	53	42	27	26	25	10	100	100	80	78	49	39	20	8	18	7
	June	100	100	56	55	53	42	28	25	24	9	100	100	80	80	49	40	20	7	17	7
pH	Sept	100	100	99	99	98	94	93	86	91	83	100	100	99	98	95	94	92	82	91	79
	Dec	100	100	100	99	88	87	93	86	94	85	100	100	100	103	99	103	97	93	97	89
	March	100	100	99	99	98	96	93	89	93	85	100	100	99	99	98	97	93	89	91	84
	June	100	100	99	99	98	93	94	93	91	84	100	100	100	99	100	95	95	89	95	85
Total hardness	Sept	100	100	102	97	103	92	109	92	110	89	100	100	102	98	103	93	107	91	108	90
	Dec	100	100	102	97	102	94	106	66	106	66	100	100	101	96	103	93	108	92	109	91
	March	100	100	105	99	103	96	105	93	109	93	100	100	104	98	105	95	109	95	102	95
	June	100	100	102	97	104	92	107	91	112	92	100	100	101	102	101	98	105	97	105	96
Chloride	Sept	100	100	100	78	97	62	94	47	94	45	100	100	95	73	95	43	86	32	86	27
	Dec	100	100	100	78	93	68	89	54	89	52	100	100	100	86	98	66	91	50	89	50
	March	100	100	97	84	95	61	93	48	90	48	100	100	100	90	96	71	93	59	89	52
	June	100	100	93	73	97	60	93	53	87	48	100	100	100	79	98	63	96	50	94	48

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CO <sub>2</sub>	Sept	100	100	120	83	130	67	170	33	190	20	100	100	107	86	136	72	162	31	197	21
	Dec	100	100	137	79	158	63	237	37	289	26	100	100	107	60	117	48	138	24	145	14
	March	100	100	113	81	132	71	148	58	161	36	100	100	109	97	131	66	163	31	191	16
	June	100	100	120	70	127	67	150	50	190	20	100	100	110	97	119	76	144	51	154	39
DO	Sept	100	100	80	120	60	136	24	180	-	188	100	100	78	117	67	139	17	217	-	228
	Dec	100	100	80	112	164	124	24	70	8	176	100	100	67	156	44	200	-	322	-	333
	March	100	100	80	125	50	145	30	175	10	210	100	100	84	110	53	132	21	189	-	195
	June	100	100	91	136	64	155	55	177	-	214	100	100	71	114	43	136	7	221	-	236
BOD	Sept	100	100	113	94	110	88	122	77	127	73	100	100	106	95	108	90	118	77	122	73
	Dec	100	100	115	96	119	90	131	76	135	72	100	100	108	96	110	91	116	81	120	77
	March	100	100	120	96	110	92	116	83	120	76	100	100	113	96	115	94	123	82	127	78
	June	100	100	115	95	113	90	124	80	131	75	100	100	111	95	114	91	118	80	123	77
COD	Sept	100	100	106	93	111	88	122	74	125	69	100	100	106	93	110	89	113	74	123	75
	Dec	100	100	107	94	111	89	125	77	129	73	100	100	104	71	106	69	113	60	114	57
	March	100	100	112	98	110	93	117	81	121	76	100	100	104	72	107	76	115	65	115	63
	June	100	100	106	94	111	88	121	79	128	71	100	100	105	97	107	92	116	84	121	83
MPN	Sept	100	100	88	71	82	59	65	35	47	12	100	100	92	83	88	75	67	54	46	33
	Dec	100	100	89	73	87	61	60	33	46	14	10	100	95	86	81	73	68	51	52	22
	March	100	100	89	94	76	56	67	47	57	34	100	100	91	78	88	70	67	50	49	48
	June	100	100	92	75	83	54	100	50	68	17	100	100	81	81	86	71	62	38	45	19
SPC	Sept	100	100	103	80	113	58	100	35	84	22	100	100	91	78	113	52	87	26	74	13
	Dec	100	100	102	88	100	68	85	46	69	34	100	100	113	77	127	59	109	34	91	23
	March	100	100	106	85	117	64	103	41	85	23	100	100	100	75	108	50	100	37	79	25
	June	100	100	95	84	97	65	84	38	70	24	100	100	90	55	105	40	110	35	85	15

C = Control values (without E. crassipes), T = Treated values using E. crassipes

respectively in comparison to the control values in the month of December, September and June while the value of control simultaneously increased from day zero to day 28 of experiments during each season. The value of turbidity and total solids in control as well as in the treated water showed a decreasing trend from zero day to 28 day but the decrease in domestic sewage was more pronounced than in the control. The self purification of sewage water to cause decrease in turbidity and total solids is well documented in the literature [150]. The amount of chloride ions present, indicated a decreasing trend with time in the control as well as in the treated sewage water. Nevertheless, the treated water exhibited more reduction in the chloride ion concentration as presented in Table 4.6. The effect of treatment on other physico-chemical parameters is summarized in Table 4.6. However, it was felt that the effect of treatment on these parameters is not very significant. The values of MPN and SPC also decreased significantly in comparison to the control from day 0 to day 28th. The maximum reduction in MPN and SPC values were 52% and 70% in the month of June 1990 respectively. The significant reduction in MPN and SPC values was observed in other months also.

The treatment of industrial effluent by E. crassipes also indicated a significant change in the water quality. The value of DO increased about 3.3 times in comparison to the control. The value of DO in control

decreased with time and reached to zero after 4 weeks. But the treated water showed an increase in DO with time. The maximum value of DO (333%) was observed in the month of Dec 1989 followed by the month of June 1990 (about 236%). The maximum reduction in CO<sub>2</sub> and BOD values was also found in the months of December 1989 and March 1990. The value of COD also indicated a minimum value in the month of December 1989, after 4 weeks treatment. The values of CO<sub>2</sub>, BOD and COD and other parameters observed for the control and treated effluent are presented in Table 4.6. The decrease in the values of chloride ions was about 40-50% throughout the period of study. The values of MPN and SPC were also decreased significantly in comparison to the control. The total reduction in MPN and SPC were about 30% and 60% in the months of December 1989 and March 1990, respectively.

#### 4.5.2 Stage II

At stage II, the water samples collected from the domestic sewage and industrial effluents were treated with the culture of L. minor plant species for a period of 4 weeks. The samples were taken out after each week and the physico-chemical and biological parameters were analysed. In the case of domestic sewage, the value of DO was always found to be more in comparison to the control. Whereas in the control the DO value decreased gradually with time. The DO value observed in the control after 4 weeks was found as zero, whereas the DO

Table 4.7 Observed values of various parameters for the stage II treated domestic sewage and industrial effluent

Parameters	Month	Domestic sewage										Industrial effluent									
		0 day		7 days		14 days		21 days		28 days		0 day		7 days		14 days		21 days		28 days	
		C	T	C	T	C	T	C	T	C	T	C	T	C	T	C	T	C	T	C	T
Turbidity	Sept	100	100	22	19	14	12	12	9	12	7	100	100	19	17	12	10	10	8	9	6
	Dec	100	100	26	22	17	13	15	11	13	6	100	100	19	17	12	10	10	8	10	6
	March	100	100	23	19	14	11	12	9	11	7	100	100	20	18	13	11	11	8	10	7
	June	100	100	24	20	5	11	13	11	12	7	100	100	20	17	12	9	10	8	10	6
Total solids	Sept	100	100	57	56	28	25	25	9	23	9	100	100	72	69	22	18	19	6	16	6
	Dec	100	100	66	64	26	24	25	9	22	9	100	100	83	69	23	20	20	7	17	7
	March	100	100	57	55	27	26	25	10	24	9	100	100	79	68	21	19	19	7	17	7
	June	100	100	54	56	28	26	25	10	24	9	100	100	80	68	23	20	20	7	17	7
pH	Sept	100	100	99	98	94	88	93	80	91	80	100	100	99	93	94	87	92	82	91	81
	Dec	100	100	100	98	95	90	94	85	94	83	100	100	107	100	99	93	97	89	97	86
	March	100	100	99	96	94	87	93	80	92	82	100	100	99	95	95	88	93	79	91	78
	June	100	100	99	96	94	86	93	80	91	80	100	100	100	100	97	90	95	85	95	83
Total hardness	Sept	100	100	102	100	105	98	109	98	110	97	100	100	102	99	105	96	107	94	107	94
	Dec	100	100	102	99	104	97	106	97	106	97	100	100	101	99	104	97	108	97	108	97
	March	100	100	102	100	105	98	109	97	110	97	100	100	102	99	105	97	108	97	108	97
	June	100	100	107	100	109	98	114	98	114	97	100	100	101	99	103	98	105	97	105	97
Chloride	Sept	100	100	100	84	97	50	96	48	95	47	100	100	95	80	93	39	91	36	91	34
	Dec	100	100	100	82	95	55	95	54	93	50	100	100	98	91	96	50	95	56	94	50
	March	100	100	97	85	95	52	94	48	94	45	100	100	100	90	98	57	97	57	93	55
	June	100	100	100	88	99	57	97	53	97	47	100	100	98	85	98	52	93	46	92	44

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CO <sub>2</sub>	Sept	100	100	120	90	153	60	170	37	190	27	100	100	107	96	145	69	155	55	189	31
	Dec	100	100	137	111	221	79	237	58	289	42	100	100	107	91	121	64	138	50	145	38
	March	100	100	117	90	153	63	167	37	187	27	100	100	113	97	155	64	168	48	187	32
	June	100	100	117	87	150	63	170	37	190	25	100	100	116	95	134	61	144	49	154	37
DO	Sept	100	100	80	92	44	52	24	36	-	16	100	100	78	100	50	61	17	72	-	44
	Dec	100	100	80	96	44	72	24	52	8	36	100	100	67	100	22	44	-	22	-	-
	March	100	100	80	95	30	45	5	30	-	15	100	100	79	95	42	53	21	37	-	21
	June	100	100	73	86	27	36	5	27	-	14	100	100	71	100	29	50	7	28	-	14
BOD	Sept	100	100	113	105	120	115	123	117	127	123	100	100	106	103	112	109	118	110	122	116
	Dec	100	100	115	107	124	118	131	121	135	129	100	100	108	105	112	110	116	112	121	115
	March	100	100	110	106	116	115	120	116	123	122	100	100	112	105	115	114	136	112	124	120
	June	100	100	122	117	129	125	134	128	141	132	100	100	111	104	116	112	119	114	123	117
COD	Sept	100	100	106	106	118	113	122	118	125	122	100	100	105	105	114	109	117	114	122	116
	Dec	100	100	107	106	119	114	125	119	118	113	100	100	104	104	110	108	113	110	114	112
	March	100	100	105	104	117	113	121	116	124	119	100	100	104	103	111	108	114	111	115	113
	June	100	100	115	114	127	125	133	126	136	130	100	100	103	103	111	108	115	111	119	114
MPN	Sept	100	100	88	82	76	53	65	47	48	35	100	100	94	92	79	77	67	60	46	38
	Dec	100	100	87	87	76	60	60	43	47	24	100	100	95	86	85	70	68	55	52	27
	March	100	100	83	78	68	51	57	41	45	28	100	100	91	87	84	71	67	28	48	16
	June	100	100	125	117	102	76	85	62	68	42	100	100	88	84	79	63	60	47	41	23
SPC	Sept	100	100	91	91	104	52	87	35	74	20	100	100	103	103	106	71	100	58	84	42
	Dec	100	100	100	100	108	60	96	42	80	20	100	100	102	102	97	76	85	61	70	51
	March	100	100	96	96	92	46	79	33	58	13	100	100	106	106	112	65	103	50	85	29
	June	100	100	115	110	110	55	95	40	70	15	100	100	95	95	95	68	84	51	70	38

C= Control values (without L. minor), T = Treatment with L. minor

level in the treated samples was found in the range of 14-36 percent (Table 4.7) during different months of investigations. The reductions in  $\text{CO}_2$ , chloride ions and SPC was more significant than other physico-chemical parameters in comparison to control. The  $\text{CO}_2$  content in the control increased with time and became 289% of zero day observation in 4 weeks. Whereas in the treated samples the  $\text{CO}_2$  content decreased with time and became 25-40 percent of the zero day values. Thus, the value of  $\text{CO}_2$  reduced to about 12-15 percent of the control after 4 weeks of treatment. The reduction in total hardness, BOD, COD, MPN, turbidity and total solids were also observed after a treatment of 4 weeks throughout the course of investigations. The details of the values observed for all the parameters during the course of investigations are presented in Table 4.7.

The analysis of industrial effluents treated with L. minor indicated that the DO value did not change for the treated samples for the first week, whereas, after 4 weeks it changed to 44 percent. The corresponding control indicated practically zero values for the DO. The maximum values of DO was observed in the month of September 1989. The trends of reduction of BOD and COD were found almost similar. The changes in BOD and COD level in the control samples were practically similar to the treated samples. Hence, it appears that BOD and COD are not affected by the treatment of effluent with L.

minor. Total hardness and pH also did not indicate a significant change by the treatment. However, the reduction in carbon dioxide value was found significant after treatment and was in the range of 25-28 percent in comparison to the corresponding values for the control.

#### 4.5.3 Stage III

The treatment of waste water with Ceratophyllum demorsum caused a change in all the physico-chemical and biological parameters, except COD, chloride ions and pH. In the case of domestic sewage, the control indicated a systematic change in DO with increasing duration. The DO decreased to about 10-25 percent after 4 weeks. The DO level for the treated samples also indicated a decreasing trend. However, the analysis at each week indicated a significant increase in DO corresponding to the control at the same time. Thus, it was concluded that C. demorsum prevented a decrease in DO of samples which would have occurred more drastically without the treatment, as indicated by the controls. The BOD value did not decrease rapidly in comparison to control. It was found that the treatment caused an insignificant change in the BOD values. The reduction of CO<sub>2</sub> was also found much lower in comparison to E. crassipes (Stage I) and L. minor (Stage II) processes. The values of MPN and SPC also found to be decreased after treatment throughout the period of investigations. The details of changes in the value of physico-chemical and biological

Table 4.8 Observed values of various parameters for the stage III treated domestic sewage and industrial effluent

Parameters	Month	Industrial effluent										Domestic sewage									
		0 day		7 days		14 days		21 days		28 days		0 day		7 days		14 days		21 days		28 days	
		C	T	C	T	C	T	C	T	C	T	C	T	C	T	C	T	C	T	C	T
Turbidity	Sept	100	100	19	17	15	14	12	11	10	10	100	100	34	30	28	24	22	19	19	15
	Dec	100	100	19	17	15	13	12	10	10	9	100	100	46	40	38	33	30	26	26	23
	March	100	100	20	18	16	14	12	13	11	10	100	100	36	31	29	25	22	18	19	15
	June	100	100	19	17	15	13	12	10	10	8	10	100	23	20	18	16	14	11	12	9
Total Solids	Sept	100	100	72	73	45	37	22	18	19	13	100	100	56	56	53	44	28	26	25	18
	Dec	100	100	82	81	50	41	23	21	20	14	100	100	66	67	56	49	26	25	24	18
	March	100	100	79	78	48	39	21	21	19	14	100	100	54	55	53	44	27	27	25	18
	June	100	100	80	80	49	40	23	21	20	14	100	100	55	54	53	44	27	27	25	18
pH	Sept	100	100	99	99	95	94	94	94	91	88	100	100	99	99	97	93	94	91	93	86
	Dec	100	100	100	100	99	98	98	95	97	91	100	100	100	98	98	94	95	93	94	88
	March	100	100	99	98	98	95	93	95	93	87	100	100	99	99	98	94	94	93	93	86
	June	100	100	100	98	100	93	97	92	95	89	100	100	99	99	98	94	94	93	93	86
Total hardness	Sept	100	100	76	74	77	73	78	72	80	72	100	100	102	100	103	99	105	99	109	98
	Dec	100	100	101	100	103	99	104	103	108	97	100	100	102	99	102	99	103	97	106	75
	March	100	100	102	99	104	99	105	97	107	97	100	100	102	100	103	99	105	99	110	98
	June	100	100	101	99	102	98	103	98	105	98	100	100	106	105	106	103	110	102	112	101
Chloride	Sept	100	100	95	89	93	86	93	77	91	77	100	100	100	98	98	98	94	94	94	92
	Dec	100	100	99	96	97	96	96	87	95	87	100	100	100	93	98	93	96	87	96	86
	March	100	100	98	95	97	93	96	86	95	86	100	100	98	97	98	93	97	87	97	87
	June	100	100	100	96	98	96	97	94	96	94	100	100	102	102	101	97	100	93	103	90

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CO <sub>2</sub>	Sept	100	100	107	107	134	96	159	90	162	76	100	100	117	107	130	100	153	97	170	76
	Dec	100	100	107	95	117	93	121	92	138	74	100	100	137	110	158	100	221	100	236	73
	March	100	100	113	103	132	100	155	97	168	84	100	100	113	103	132	97	145	93	161	84
	June	100	100	107	102	117	97	131	96	140	90	100	100	120	107	140	100	150	96	167	90
DO	Sept	100	100	78	83	66	78	50	61	16	44	100	100	80	92	60	72	44	52	24	36
	dec	100	100	66	100	44	66	22	44	-	33	100	100	80	96	64	84	44	72	24	52
	March	100	100	79	95	56	79	42	52	21	37	100	100	80	95	50	65	30	45	10	30
	June	100	100	71	100	43	71	29	50	7	28	100	100	73	86	45	59	27	51	9	23
BOD	Sept	100	100	106	105	108	106	111	109	117	110	100	100	122	110	118	118	134	126	138	129
	Dec	100	100	108	104	110	108	112	110	116	109	100	100	128	116	138	122	147	135	160	142
	March	100	100	112	105	113	108	114	114	121	115	100	100	118	110	118	116	127	124	134	126
	June	100	100	111	104	114	110	116	112	118	113	100	100	122	115	120	120	127	124	130	125
COD	Sept	100	100	106	105	110	109	114	114	118	117	100	100	106	105	110	110	118	115	122	120
	Dec	100	100	104	104	106	106	110	109	113	110	100	100	107	102	111	108	119	114	125	119
	March	100	100	103	103	105	105	111	111	115	114	100	100	104	104	110	108	116	113	121	117
	June	100	100	103	103	106	104	111	107	114	110	100	100	115	112	120	117	127	120	131	129
MPN	Sept	100	100	96	82	88	83	79	75	66	62	100	100	88	82	82	59	76	58	65	47
	Dec	100	100	95	95	90	81	84	68	68	75	100	100	89	86	86	70	76	63	60	43
	March	100	100	91	89	88	78	69	83	67	52	100	100	83	83	75	64	68	55	56	42
	June	100	100	99	85	88	77	86	64	79	51	100	100	125	121	113	96	101	83	86	60
SPC	Sept	100	100	103	103	113	90	106	74	100	61	100	100	91	91	113	82	104	52	87	43
	Dec	100	100	102	102	100	88	97	80	85	63	100	100	112	100	108	76	96	64	96	40
	March	100	100	106	105	117	88	114	111	103	50	100	100	96	95	108	66	95	50	79	33
	June	100	100	96	94	97	84	94	70	84	54	100	100	126	115	130	80	110	60	95	40

C = Control values ( without C. demorsum ), T = Treatment with C. demorsum

parameters of control as well as treated samples, with time (from zero day to 4 weeks) are presented in Table 4.8.

The treatment of industrial effluents with C. demorsum at stage III also indicated a behaviour similar to the domestic sewage.

A comparison of values of various parameters determined for sewage and industrial effluents in the month of September 1989 is presented in Table 4.9. It is clearly seen that the values observed for domestic sewage after 4 weeks treatment provide a water of reasonably pure quality, the values observed are well within the range suggested by WHO for agriculture purposes [161]. On the other hand the value of DO in the case of sewage as well as effluents is very low and hence even after the treatment the water is not suitable for bathing or drinking purposes. Hence, it could be concluded that the treated water for municipal sewage as well as effluent is not reasonably pure even for bathing due to high values of MPN and SPC and therefore some other treatments are necessary before it is mixed in the river Dhamola.

#### 4.5.4 Stage IV

As water hyacinth (E. crassipes, Stage I) was found more effective in reducing BOD, COD, MPN and CO<sub>2</sub> of the waste water samples and L. minor (Stage II) was

Table 4.9 A comparison of various physico-chemical and biological parameters for domestic sewage and industrial effluent after the treatment with C. demorsum for a period of 4 weeks in the month of September, 1989.

Parameters*	Domestic sewage	Industrial effluent
Turbidity (NTU)	27.5	36.0
Total Solids (mg/l)	129.0	146.0
pH	7.0	7.5
Total Hardness (mg/l)	275.0	284.0
Chloride (mg/l)	29.6	17.0
DO (mg/l)	0.9	0.8
BOD (mg/l)	181.0	196.0
COD (mg/l)	294.0	245.0
CO <sub>2</sub> -free (mg/l)	2.3	2.2
MPN x 10 <sup>4</sup> (per 100 ml)	9.9	17.0
SPC x 10 <sup>5</sup> (per ml)	12.0	23.0

\* All values reported are an average of atleast two replicate determinations.

useful in reducing turbidity, total solids, chlorides and total hardness, it was considered interesting to observe the effect of equal amount of both the plant species together. In stage IV, a mixture of both the plant species (E. crassipes and L. minor) was taken in equal amount for the treatment of waste water. It was observed that the concentration of DO in stage IV increased much more than in the case of individual plant species during the entire course of investigation. The values of DO for the controls and treated samples are presented in Table 4.10. The BOD and COD values also decreased to about 50% of the control values during the entire period. Turbidity, total solids, chlorides and CO<sub>2</sub> indicated a significant decrease in the sewage water after the treatment at stage IV. Thus, it seems reasonable to conclude that the mixture of two different plant species, viz., water hyacinth (E. crassipes) and L. minor are more effective in removing pollution load in sewage water in comparison to the individual species.

The industrial effluent treated at stage IV also indicated a similar behaviour in which DO increased to a level similar to the domestic sewage. BOD, COD and other parameters indicated a decrease in the values. However, an overall effect of the two species on physico-chemical and biological parameters for the sewage water and effluent was found practically similar and are summarized in Table 4.10.



Table 4.10 Observed values of various parameters for the stage IV treated domestic sewage and industrial effluent

Parameters	Month	Industrial effluent										Domestic sewage									
		0 day		7 days		14 days		21 days		28 days		0 day		7 days		14 days		21 days		28 days	
		C	T	C	T	C	T	C	T	C	T	C	T	C	T	C	T	C	T	C	T
Turbidity	Sept	100	100	19	14	15	10	10	5	9	4	100	100	22	16	14	7	12	6	11	4
	Dec	100	100	19	14	12	6	10	5	10	4	100	100	26	20	17	9	15	7	13	6
	March	100	100	20	15	13	7	11	6	10	4	100	100	23	16	14	6	12	6	11	4
	June	100	100	19	14	12	6	10	5	9	4	100	100	21	15	14	7	12	5	11	4
Total solids	Sept	100	100	72	68	45	33	19	5	16	5	100	100	56	55	28	25	25	7	23	7
	Dec	100	100	82	66	23	20	20	6	17	5	100	100	65	55	26	24	25	7	22	6
	March	100	100	79	64	20	19	19	5	17	5	100	100	54	42	27	25	24	7	23	6
	June	100	100	80	62	23	19	20	6	17	5	100	100	56	56	28	25	24	7	23	6
pH	Sept	100	100	99	99	95	93	92	82	91	77	100	100	99	109	94	89	92	84	91	80
	Dec	100	100	100	102	98	89	97	97	97	85	100	100	100	99	95	94	94	85	94	80
	March	100	100	99	100	95	93	93	85	91	79	100	100	99	98	94	89	92	84	91	80
	June	100	100	100	97	97	87	95	97	95	81	100	100	99	96	94	87	92	81	91	77
Total hardness	Sept	100	100	102	97	103	92	107	91	108	90	100	100	102	95	105	90	109	89	110	89
	Dec	100	100	101	97	104	92	108	92	109	91	100	100	102	99	104	94	106	94	106	93
	March	100	100	102	97	105	92	107	91	108	91	100	100	102	95	105	91	109	91	110	89
	June	100	100	101	103	103	99	105	98	105	98	100	100	102	95	106	92	111	92	112	90
Chloride	Sept	100	100	95	66	95	41	91	29	91	27	100	100	100	78	97	50	97	45	96	44
	Dec	100	100	100	82	98	50	97	55	96	50	100	100	100	78	98	53	96	53	95	52
	March	100	100	100	91	98	55	97	53	96	53	100	100	97	74	96	47	92	45	90	42
	June	100	100	100	77	98	52	95	50	94	46	100	100	100	80	103	53	103	45	103	47

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CO <sub>2</sub>	Sept	100	100	107	86	134	69	162	31	195	17	100	100	117	81	150	47	170	30	188	17
	Dec	100	100	88	71	121	43	138	57	145	21	100	100	189	84	221	53	231	37	281	26
	March	100	100	113	93	155	58	168	29	187	13	100	100	113	90	145	55	161	32	181	13
	June	100	100	107	93	131	62	140	45	150	33	100	100	119	73	147	47	170	30	189	17
DO	Sept	100	100	78	111	67	133	17	217	-	222	100	100	80	125	44	152	24	176	-	184
	Dec	100	100	67	155	22	322	-	266	-	333	100	100	80	108	44	140	24	160	8	172
	March	100	100	79	105	42	163	21	184	-	189	100	100	80	130	30	170	10	210	-	215
	June	100	100	71	107	28	185	7	221	-	228	100	100	91	136	45	173	27	200	-	214
BOD	Sept	100	100	106	94	108	89	117	74	122	72	100	100	113	93	120	81	122	76	127	72
	Dec	100	100	108	94	112	80	116	84	120	76	100	100	115	95	124	83	131	74	135	71
	March	100	100	112	93	115	83	121	79	124	76	100	100	111	92	116	82	120	75	123	72
	June	100	100	111	95	116	84	118	78	123	76	100	100	115	95	123	84	127	78	131	73
COD	Sept	100	100	106	93	110	88	117	77	122	74	100	100	106	93	118	80	122	73	125	69
	Dec	100	100	104	70	110	60	113	64	114	57	100	100	107	94	119	81	125	77	129	73
	March	100	100	104	70	111	70	114	64	114	61	100	100	104	93	117	79	121	75	124	71
	June	100	100	103	69	111	86	115	82	119	80	100	100	107	92	120	81	125	74	128	69
MPN	Sept	100	100	94	79	88	71	67	50	46	30	100	100	88	70	76	44	65	32	48	11
	Dec	100	100	95	86	84	48	68	59	52	20	100	100	89	70	76	53	60	30	47	12
	March	100	100	91	73	83	63	67	48	48	24	100	100	83	67	68	44	57	32	45	20
	June	100	100	88	74	79	53	60	32	44	18	100	100	117	83	108	58	92	42	67	16
SPC	Sept	100	100	103	77	113	55	100	32	86	22	100	100	91	78	104	39	87	26	74	130
	Dec	100	100	102	85	97	44	85	56	69	32	100	100	100	68	108	34	96	28	80	20
	March	100	100	106	76	111	50	103	30	85	20	100	100	96	75	92	37	79	25	58	12
	June	100	100	95	81	95	49	84	35	70	22	100	100	100	85	120	40	100	31	85	15

C = Control values ( without E. crassipes and L. minor ), T = Treatment with E. crassipes and L. minor

#### 4.5.5 Stage V

At stage V, the waste water samples were treated with a mixture of L. minor and C. demorsum plant species. It was observed that the mixture of these plant species had almost similar effect on the physico-chemical and biological parameters of domestic sewage and effluents as was observed in the case of individual species. Thus the turbidity, total solids, total hardness, chloride ions, CO<sub>2</sub>, BOD, COD and SPC reduced to an extent as was observed in stage II and stage III treatment separately. The value of DO was increased to about 25-30 percent in the presence of L. minor and C. demorsum in comparison to the control after 4 weeks. In the individual plants treatment at stage II and III, the maximum increase in DO was 10-20 percent. The details of all the physico-chemical and biological parameters are presented in Table 4.11.

In general, Eichhornia crassipes (Stage I) has been found more effective in reducing the pollution load in the present studies. It may be due to the fact that E. crassipes is a large leaf plant. As large leaf covers large area, more transpiration occurs during the photosynthesis process which leads the increase in DO level. Thus, increase in DO level at stages I and IV may be due to the large leaf area, well developed arenchyma and root system of E. crassipes. However, decrease in BOD and COD values may be due to better availability of

**Table 4.11** Observed values of various parameters for the stage V treated domestic sewage and industrial effluent

Parameters	Month	Domestic sewage										Industrial effluent									
		0 day		7 days		14 days		21 days		28 days		0 day		7 days		14 days		21 days		28 days	
		C	T	C	T	C	T	C	T	C	T	C	T	C	T	C	T	C	T	C	T
Turbidity	Sept	100	100	19	16	12	8	10	6	9	5	100	100	22	18	14	9	12	7	11	5
	Dec	100	100	19	15	12	8	10	7	10	6	100	100	26	21	17	11	15	9	13	6
	March	100	100	20	16	13	8	11	7	10	6	100	100	23	18	14	8	12	7	11	5
	June	100	100	19	15	11	8	10	6	9	5	100	100	21	18	13	13	11	7	11	5
Total solids	Sept	100	100	72	70	21	18	19	7	16	7	100	100	56	55	27	25	25	10	23	10
	Dec	100	100	82	68	22	20	20	8	17	8	100	100	65	63	26	25	25	10	22	9
	March	100	100	78	68	20	18	19	8	16	8	100	100	57	54	27	26	25	11	23	10
	June	100	100	80	69	23	19	20	9	17	7	100	100	56	55	27	24	24	10	22	10
pH	Sept	100	100	99	95	94	89	92	83	90	82	100	100	99	97	94	87	92	85	91	80
	Dec	100	100	100	105	98	94	97	89	97	89	100	100	100	100	95	92	94	87	93	86
	March	100	100	100	99	96	95	93	83	91	83	100	100	98	97	94	87	92	84	91	80
	June	100	100	100	101	97	92	95	87	95	85	100	100	98	97	94	96	91	84	91	80
Total hardness	Sept	100	100	102	98	105	95	107	91	107	90	100	100	102	97	104	95	109	95	109	90
	Dec	100	100	103	96	104	91	108	91	108	90	100	100	101	98	104	92	106	92	106	65
	March	100	100	102	98	104	94	107	94	108	94	100	100	102	98	105	92	109	92	110	101
	June	100	100	103	101	103	96	105	96	105	95	100	100	73	98	106	89	109	91	109	91
Chloride	Sept	100	100	95	68	91	32	91	32	88	27	100	100	100	78	97	50	95	47	95	45
	Dec	100	100	100	87	98	55	96	50	96	50	100	100	100	78	97	53	97	53	95	50
	March	100	100	100	89	98	55	96	53	94	52	100	100	97	79	95	47	95	46	93	45
	June	100	100	100	77	98	50	98	48	96	46	100	100	103	73	100	50	100	46	98	46

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CO <sub>2</sub>	Sept	100	100	107	93	159	69	162	55	193	31	100	100	120	90	153	60	170	37	190	23	
	Dec	100	100	107	88	121	62	136	48	140	36	100	100	137	105	221	79	237	53	289	37	
	March	100	100	113	97	155	61	168	48	190	29	100	100									
	June	100	100	107	93	131	57	141	45	150	36	100	100	100	87	147	57	67	37	183	27	
DO	Sept	100	100	78	105	50	67	17	50	-	33	100	100	80	92	44	60	24	40	-	24	
	Dec	100	100	67	111	22	67	-	44	-	11	100	100	80	100	44	76	24	40	8	24	
	March	100	100	79	108	42	68	21	42	-	26	100	100	50	100	40	50	5	37	-	20	
	June	100	100	71	107	29	64	7	43	-	14	100	100	91	100	45	64	27	41	-	27	
BOD	Sept	100	100	106	101	111	108	117	109	122	116	100	100	113	104	120	112	122	116	127	122	
	Dec	100	100	108	103	112	109	116	111	120	114	100	100	115	106	124	116	131	119	135	127	
	March	100	100	64	60	66	65	70	66	71	68	100	100	111	104	116	112	120	115	123	120	
	June	100	100	111	102	116	108	119	113	123	115	100	100	116	108	108	116	123	116	129	125	
COD	Sept	100	100	106	101	114	108	117	114	122	87	100	100	106	102	118	111	122	116	125	119	
	Dec	100	100	104	101	110	106	113	109	114	111	100	100	107	103	119	112	125	117	129	121	
	March	100	100	104	103	111	106	114	109	114	111	100	100	104	104	117	110	121	114	123	116	
	June	100	100	103	102	111	106	115	109	119	111	100	100	108	102	120	113	121	117	126	121	
MPN	Sept	100	100	94	87	79	71	67	58	46	35	100	100	88	82	76	53	65	41	48	14	
	Dec	100	100	95	84	84	70	68	52	52	25	100	100	89	80	76	53	60	40	47	20	
	March	100	100	91	83	84	69	67	52	48	30	100	100	83	72	68	50	57	39	45	27	
	June	100	100	88	79	79	60	60	46	44	21	100	100	125	116	108	75	92	58	68	21	
SPC	Sept	100	100	103	100	106	68	100	58	84	39	100	100	91	83	104	52	87	39	74	19	
	Dec	100	100	102	97	97	76	85	61	69	51	100	100	100	88	108	56	96	40	80	20	
	March	100	100	106	100	112	62	103	47	85	26	100	100	96	87	92	48	79	33	58	12	
	June	100	100	95	89	95	67	84	49	70	35	100	100	105	95	120	60	100	45	85	22	

C = Control values ( without L. minor & C. demorsum ), T = Treatment with L. minor & C. demorsum

oxygen in root tissue of E. crassipes. The change in the pH value during the treatment process is possibly due to the volatilization as suggested by Stratton [170]. The significant decrease in MPN and SPC (coliform bacteria) by E. crassipes is well reported in the literature [169] and has been assigned to the agglomeration of colloidal particles in the waste water which carries the bacteria to the bottom [171].

The practically similar values of BOD and COD of the control and treated samples at stage II (L. minor) can be assigned to the accumulation of the decayed plant at the bottom of the containers during the treatment [4]. Such an accumulation would result in the degradation of organic matter by micro-organisms. These organisms consume the dissolved oxygen for the decomposition of organic matter and cause an increase in BOD and COD values.

#### 4.5.6 Biomass of Hydrophytes

The percentage changes in the biomass of different hydrophytes after a treatment of four weeks for domestic waste and industrial effluents were carried out for all the plant species used in the investigation. The amount of biomass obtained after a 4 weeks treatment in various months are depicted in fig. 4.8. It can be clearly seen that the biomass of E. crassipes is almost 2 to 3 times greater than all the other plant species used in the

domestic sewage as well as in industrial effluents. L. minor and C. demorsum on the other hand indicated a loss in biomass. Such a behaviour of L. minor and C. demorsum is well reported in the literature [172] and has been accounted for on the basis of destruction of the membrane system of the chloroplast of these hydrophytes in the waste water. Huber et al. [173] have also found a reduction in biomass of these hydrophyte due to the reduction of the photosynthesis of aquatic macrophytes by the waste water.

Thus, it is clear from the treatment of domestic sewage and industrial effluents by various plant species that these plants are capable of reducing the organic and inorganic pollution load of waste water. However, a comparison of various treatment clearly indicated that such a treatment alone is not sufficient for making the water useful for bathing or drinking purposes. The analysis of the various parameters pointed out that the treatment with the plant species reduces the pollution load to such an extent so that the water samples can be used for agricultural purposes. Nevertheless, it is felt that some more treatment are necessary to make it more useful and should also be carried out before mixing it in the river Dhamola to control the pollution load in the river water.

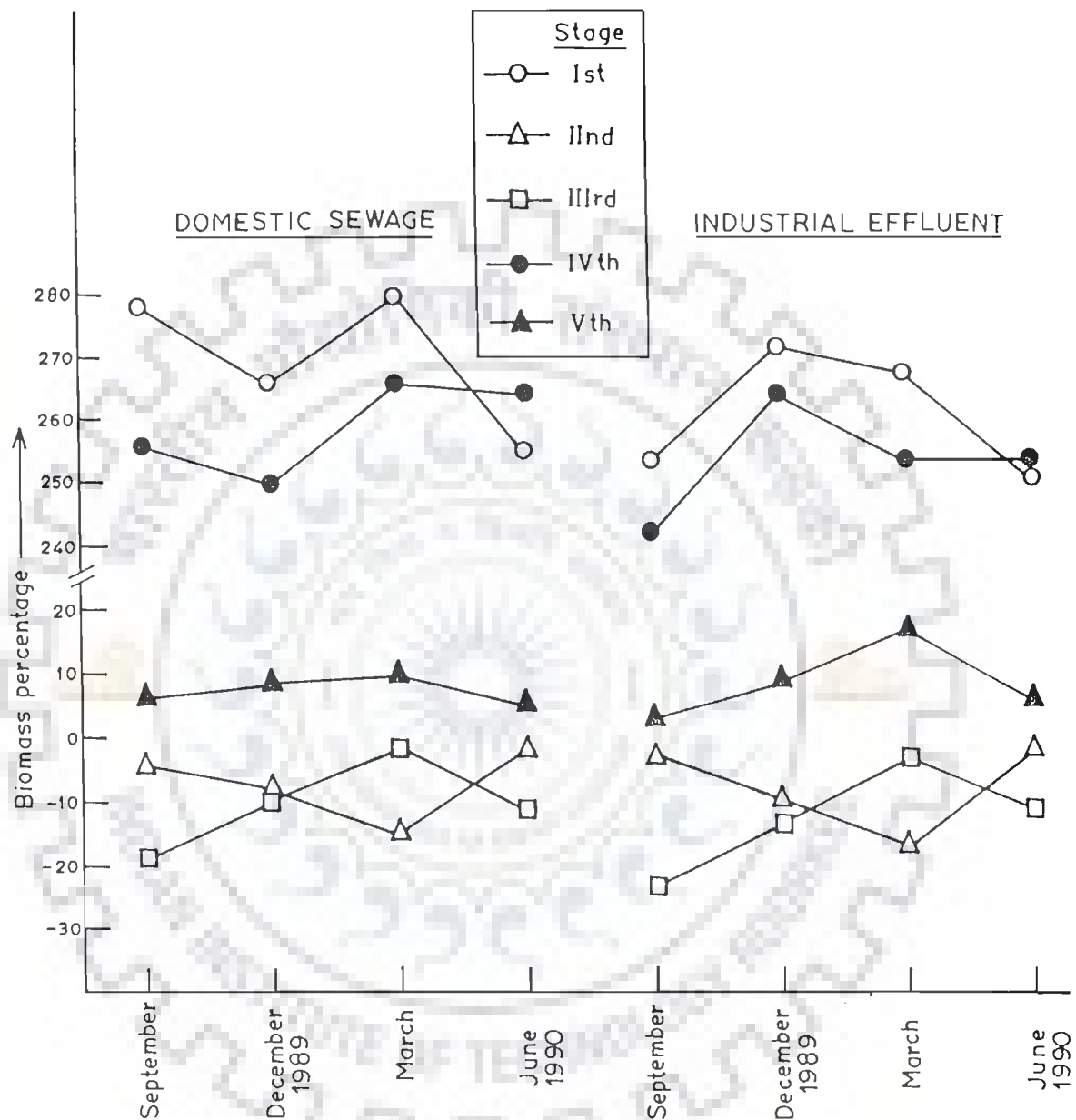


Fig. 4.8 Percentage change in biomass of treated hydrophytes after 4 weeks during different seasons of 1989-90



#### 4.6 STATISTICAL CORRELATIONS

Correlations among water quality parameters provides an important information about the overall quality of water. Since other parameters and their functions can be explained on the basis of these correlations. Attempts have also been made to calculate the correlation coefficients among the physico-chemical and biological parameters determined for the three sampling stations for the river Dhamola. The entire information collected for a period of 2 years was used for determining the value of correlation coefficient 'r' which was calculated using the equation -

$$r = \frac{\sum YX}{(\sum X^2 \cdot \sum Y^2)^{\frac{1}{2}}}$$

Where,

$$x = (x - \bar{x}), \quad y = (y - \bar{y})$$

$$\bar{x} = \sum X/n, \quad \bar{y} = \sum y/n$$

n = total number of observations.

To carry out the extensive numerical calculations in a systematic manner, a LOTOUS programme ( by using a package ) was used to calculate the parameters A and B of linear regression line,  $y = A + Bx$ . For fourteen parameters determined in the present investigation, the total numbers of correlation possible were ninety one. All numerical values of correlation coefficient 'r' for 91 pairs are presented in Table 4.12. It was observed

Table 4.12 Correlation coefficient values observed among various physico-chemical and biological parameters

	Total solids	pH	Total hardness	Chloride	CO <sub>2</sub>	DO	BOD	COD	MPN	SPC	Algae	Diatoms	Protozoans
Turbidity	0.89	0.29	0.51	0.55	0.34	0.82	0.82	0.66	0.58	0.84	0.76	0.76	0.76
Total solids		0.29	0.49	0.49	0.30	0.69	0.77	0.73	0.52	0.79	0.70	0.60	0.58
pH			0.22	0.14	0.09	0.29	0.28	0.31	0.15	0.24	0.25	0.33	0.31
Total hardness				0.33	0.22	0.64	0.64	0.63	0.41	0.63	0.60	0.45	0.42
Chloride					0.55	0.67	0.66	0.71	0.36	0.65	0.72	0.62	0.70
CO <sub>2</sub>						0.50	0.46	0.50	0.14	0.37	0.51	0.51	0.54
DO							0.93	0.96	0.46	0.83	0.94	0.81	0.80
BOD								0.97	0.55	0.93	0.95	0.85	0.81
COD									0.49	0.89	0.98	0.87	0.85
MPN										0.63	0.47	0.39	0.40
SPC											0.86	0.73	0.70
Algae												0.87	0.86
Diatoms													0.95

Table 4.13 Least square fitting of the relation  $y = A+Bx$  among various parameters

x	y	r*	A	B
Turbidity	Total solids	0.89	383.48	1.28
DO	BOD	0.93	180.75	-21.30
DO	COD	0.96	531.32	-62.47
DO	Algae	0.94	-1778.48	394.85
BOD	COD	0.97	6.40	2.84
BOD	SPC	0.93	-112640.04	8816.85
BOD	Algae	0.95	1538.52	-17.89
COD	SPC	0.89	114560.72	2994.52
COD	Algae	0.98	86.16	-6.32
Diatoms	Protozoans	0.94	57.4	1.07

\*r- Correlation coefficient between x and y

that the value of  $r$  was significant, i.e. equal or larger than 0.89 for only ten pairs of parameters. The values of  $A$  and  $B$  for these ten significant pairs were calculated and are presented in Table 4.13. A detailed description of the ten significant correlations is presented below:

The correlation between turbidity and total solids of the collected water samples for river Dhamola is found to be statistically significant. The values of  $r$ ,  $A$  and  $B$  calculated by the linear regression output were found as 0.89, 383.48 and 1.28, respectively. Thus, both the parameters are dependent on each other, i.e., if amount of total solids is increased, turbidity also increased and hence a straight line is obtained between two parameters (Fig. 4.9a). It can be concluded on the basis of the observed behaviour that the turbidity in the water of Dhamola river was mainly due to the presence of total solids present in the river water.

A graphical relationship between dissolved oxygen and biochemical oxygen demand is presented in Figure - 4.9b. It can be clearly seen that an inverse relation, i.e. increase in BOD caused decrease in DO. The value of  $r$  determined for the linear relation was found to be 0.93. The observed relationship further confirmed our earlier observation that the value of DO decreased due to the mixing of organic matter which increased the value of BOD and therefore indicated pollution in the river water [174].

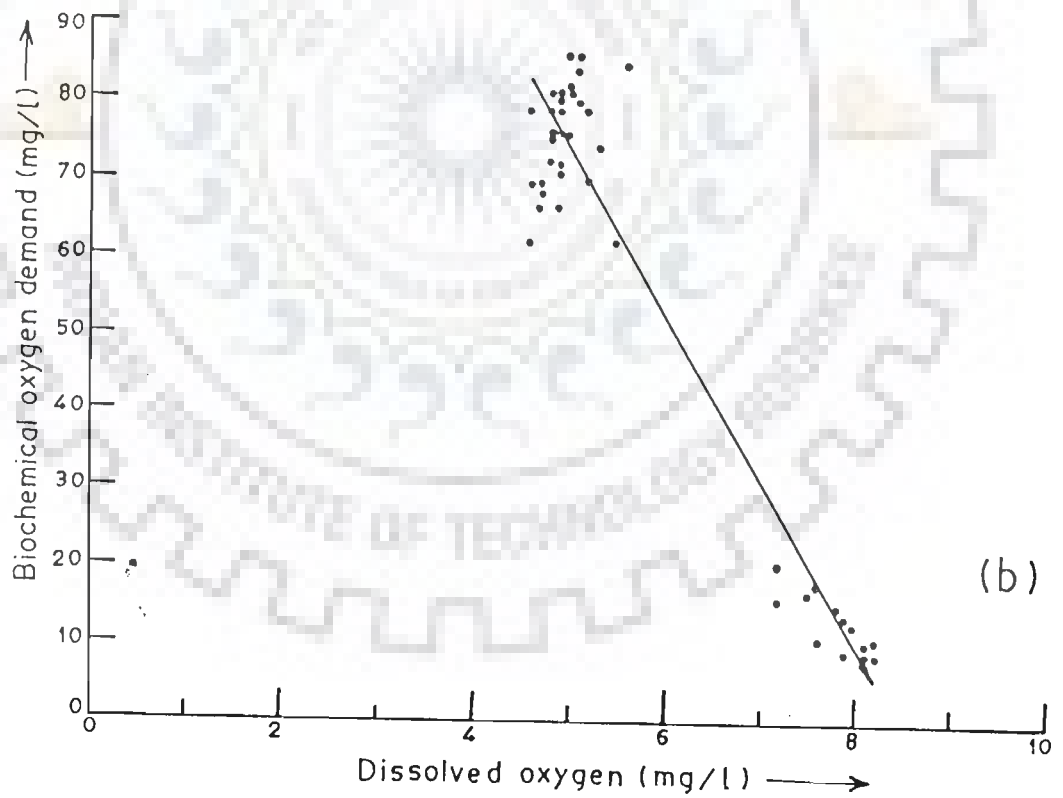
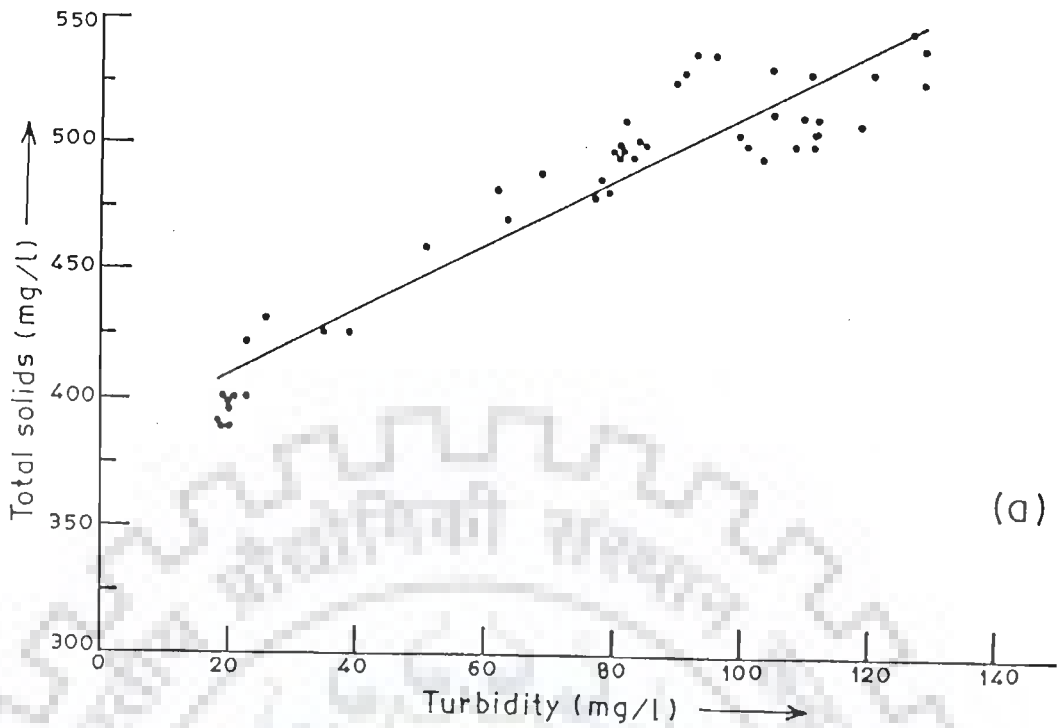


Fig. 4.9 Linear relations observed between turbidity and total solids (a) and DO and BOD (b)

The correlation between DO and COD is also found statistically significant and value of 'r' was found as 0.96. The dependence of DO on COD for various samples obtained during a period of two years is presented in Fig.4.10a and indicates an inverse relationship, which indicates that the decrease in DO will cause an increase in COD value. Such a behaviour is a common phenomenon as mixing of organic/inorganic materials in river Dhamola will require DO for chemical oxidation and hence DO will decrease [174]. Therefore, the value of COD will increase and hence an inverse relation is expected in such cases. In the present investigation similar behaviour is observed and therefore it can be safely concluded that DO is inversely dependent on the COD value obtained for river Dhamola.

The relationship between DO and algae (primary producer) is presented in Fig. 4.10b and has been found significant. The value of r, obtained for a linear regression was 0.94. This relationship indicate that DO value of water samples increased due to the increase of algal population which liberates oxygen (free as well as dissolved) in aquatic system. An increase in DO due to increase of algal population has been well documented in literature [141]. The values of regression output are summarised in Table 4.13.

The correlation obtained between BOD and COD values was found as statistically significant with the

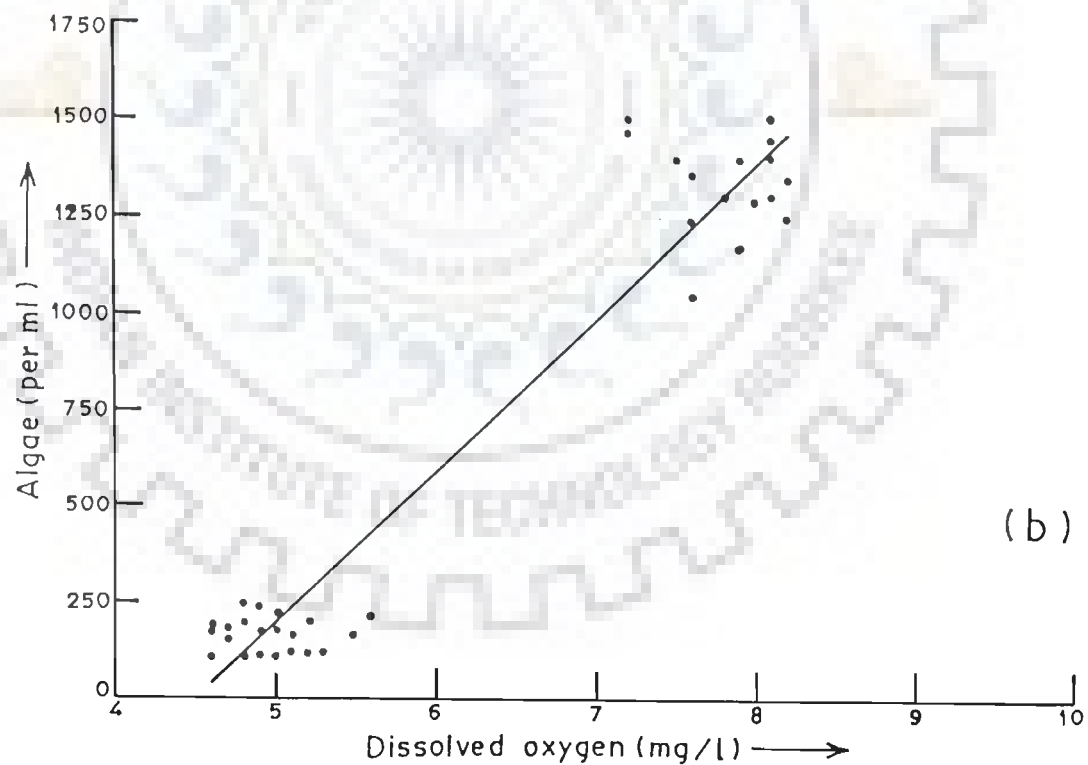
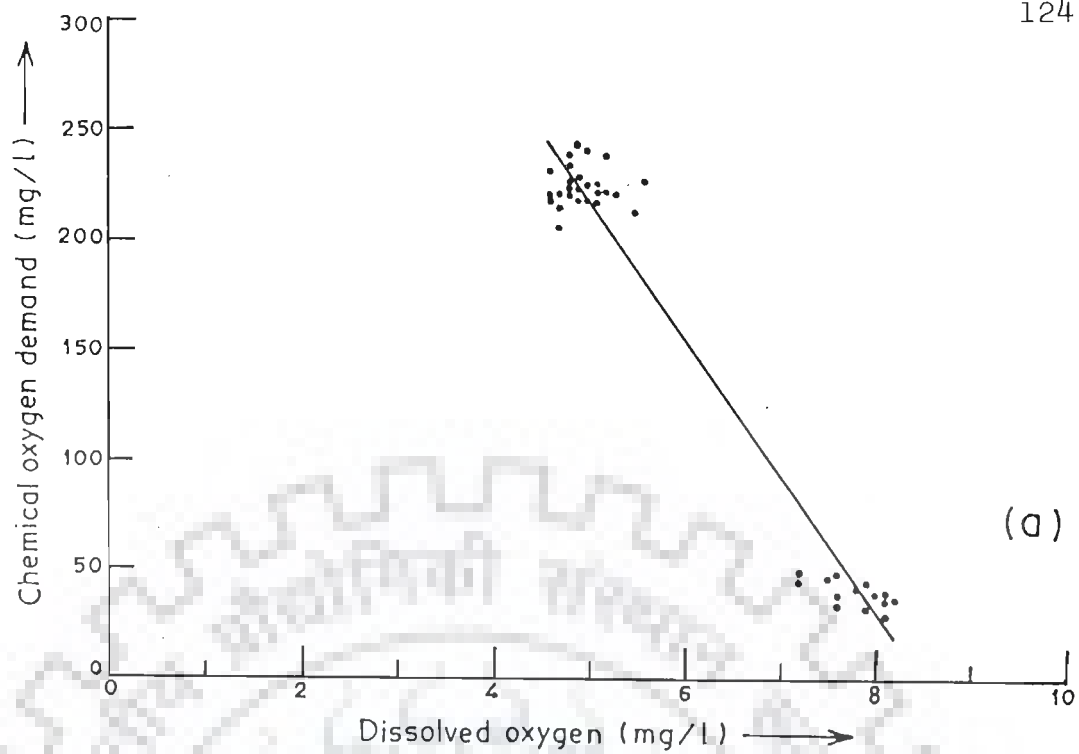


Fig. 4.10 Linear correlations observed between DO and COD (a) and DO versus algae (b)

value of  $r$  as 0.97. A direct linear relation between BOD and COD clearly indicates that an increase in BOD caused a simultaneous increase in COD of the water samples. Thus an increase in organic pollution load in river Dhamola will indicate an increase in BOD as well as COD of water samples collected at different points. The details of regression output of BOD and COD are presented in Fig.4.11a. An increase in BOD and COD due to the mixing of organic pollution load in Ganga river has been observed earlier by Tiwari et al.[175].

The relationship between BOD and SPC (micro-organisms, coliform bacteria) is also found statistically significant with  $r = 0.93$ . The values of all the regression factors are given in Table 4.13. A direct linear line between BOD and SPC (Fig.4.11b) clearly indicated that increase in BOD caused an increase in SPC. Thus, after mixing of waste water the population of microorganisms (coliform bacteria, SPC) increased tremendously, which caused an increase in BOD of the water samples. A linear relation between microorganisms and BOD has also been observed earlier by Chiu and Frost for Medlock river in England [176].

The correlation between BOD and algae on the other hand was inverse and statistically significant ( $r = 0.95$ ). The details of regression output are given in Table 4.13 and Fig. 4.12a.



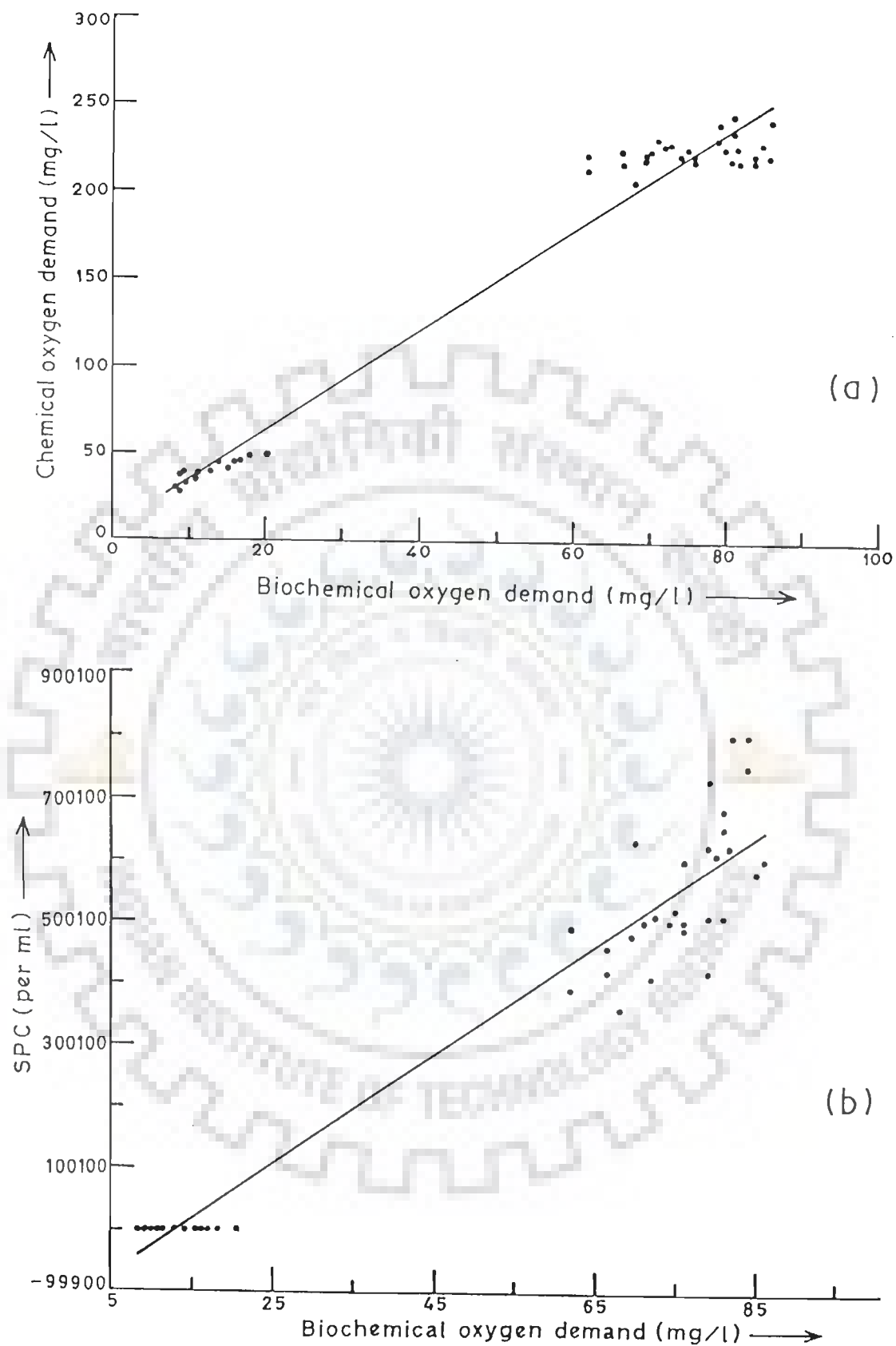


Fig. 4.11 Linear relation obtained between BOD versus COD (a) and BOD versus SPC (b)

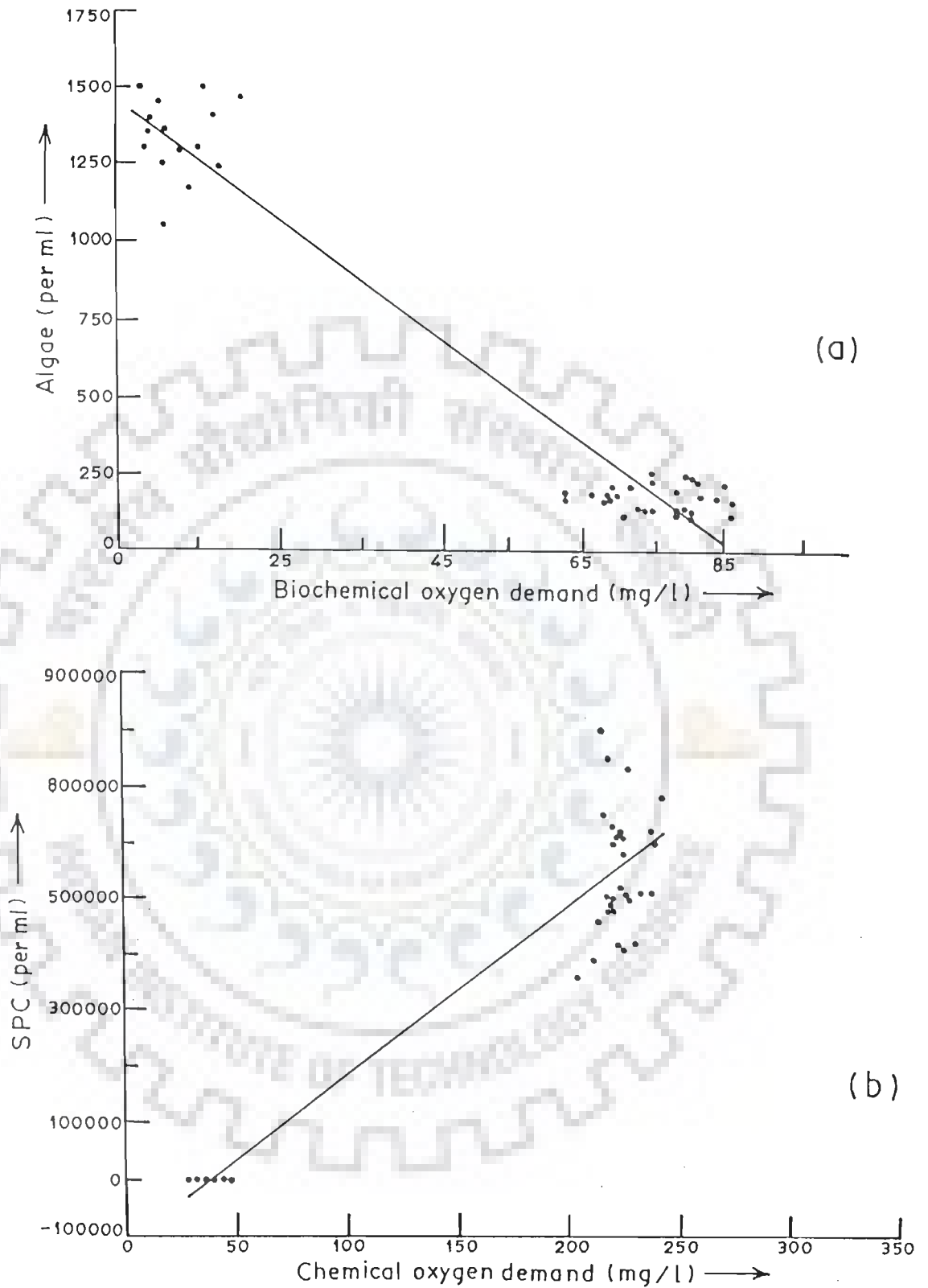


Fig. 4.12 Linear relation observed between BOD versus Algae (a) and COD versus SPC (b)

Fig.4.12 b , represents the statistical correlation between COD and SPC. The regression output between these parameters was a direct linear line. Thus COD of water samples increased due to the microorganisms (coliform bacteria, SPC). The microorganisms decompose the organic matter present in water and consume the oxygen, which causes an increase in COD value. The value of correlation coefficient (r). A and B are presented in Table 4.13.

A relationship between COD and algae is also found statistically significant and confirmed that the pollution load in river water was observed due to the high value of COD. As observed earlier, at high COD values the algal population was low. The value of correlation coefficient was 0.98. The inverse linear regression for COD and algae is presented in Fig. 4.13a

The value of 'r' for diatoms and protozoans number is significant (0.94) and directly correlates with each other. A direct linear line between diatoms and protozoans indicates that increase in diatoms in water samples collected also caused an increase in protozoans. The linear relation between diatoms and protozoans is presented in Fig. 4.13b

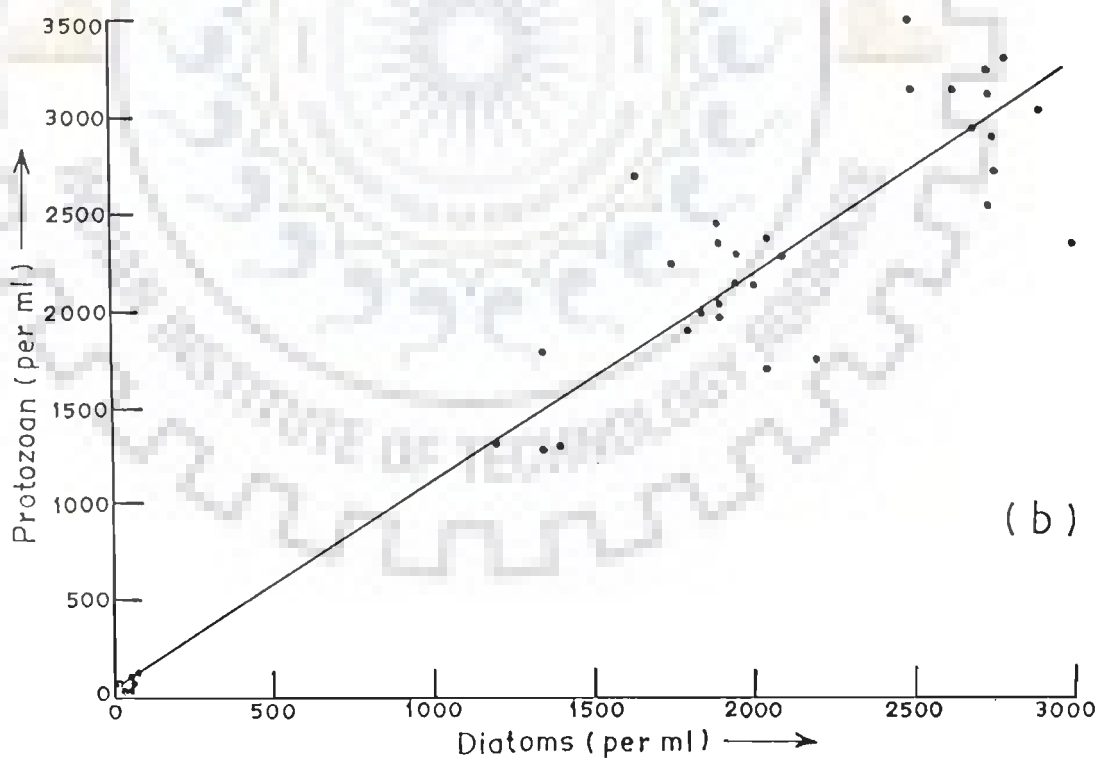
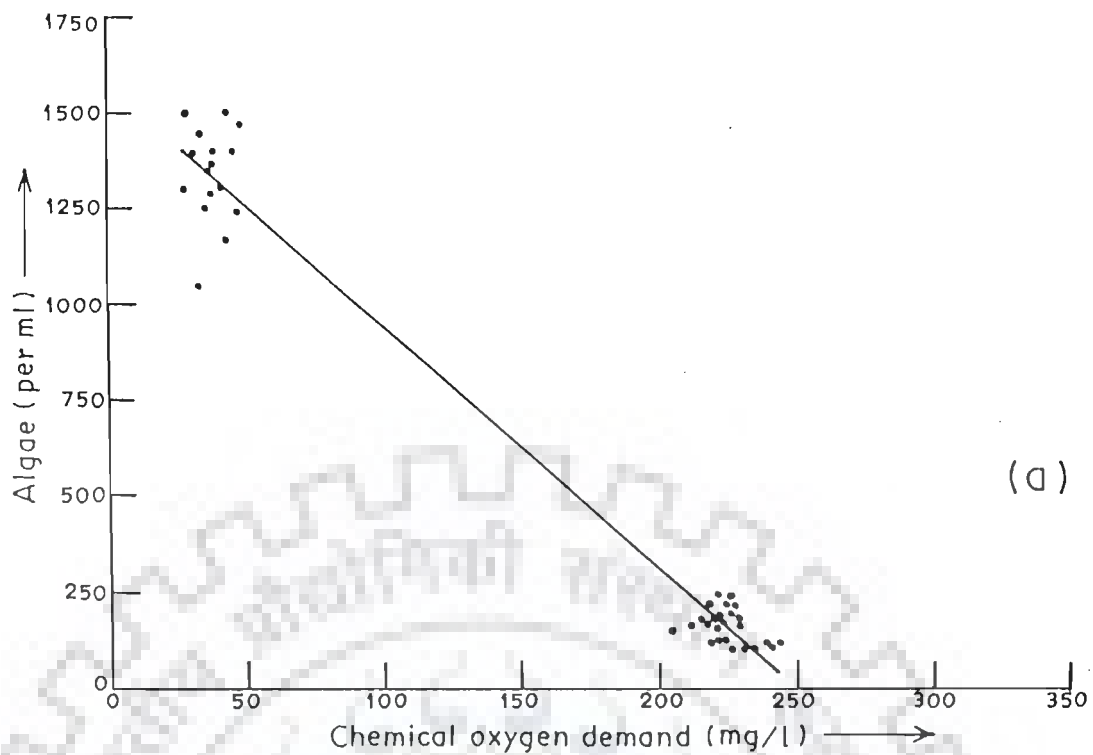


Fig. 4.13 Linear correlation observed between COD versus algae (a) and diatoms versus protozoans (b)



CHAPTER - V  
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